

Article



# **Optimum Particle Size of Treated Calcites for CO<sub>2</sub> Capture in a Power Plant**

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Abstract: This work has analyzed the influence of the particle size of a calcite from a quarry, whether original, calcined, or rehydrated, on the efficiency of CO<sub>2</sub> capture of the gases emitted in a coal-fired power plant. Three different particle sizes 0.5 mm, 0.1 mm, and 0.045 mm have been studied. The calcination had a minimal effect on the particle size of the smaller samples A1045 and A1M1 (<30  $\mu$ m). The N<sub>2</sub> isotherms and the CO<sub>2</sub> adsorption isotherms at 0 °C showed a very significant increase in the surface of the calcined and rehydrated samples (A15CH, A1045CH, and A1M1CH) with respect to the calcined or original samples. The results obtained showed that the capture of CO<sub>2</sub> for the sample A1M1, with a smaller average particle size (<30  $\mu$ m, is the most effective. For the sample A1M1 calcined and completely rehydrated (Ca(OH)<sub>2</sub>), the chemical adsorption of CO<sub>2</sub> to form CaCO<sub>3</sub> is practically total, under the experimental conditions used (550 °C and CO<sub>2</sub> flow of 20 mL min<sup>-1</sup>). The weight increase was 34.11% and the adsorption capacity was 577.00 mg g<sup>-1</sup>. The experiment was repeated 10 times with the same sample A1M1 calcined and rehydrated. No appreciable loss of adsorption capacity was observed.

**Keywords:** greenhouse gas emissions; carbon dioxide; adsorption; calcite; calcite calcined and rehydrates

# 1. Introduction

Up to 55% of greenhouse gas emissions corresponds to  $CO_2$ . Nowadays there is a growing concern in several countries regarding the impact caused by emissions of greenhouse gases such as  $CO_2$ .

The International Energy Agency (IEA) [1] forecasts that demand for energy will increase 40% by 2030, an average annual increase of 1.5% with over 50% of the increase in demand coming from China and India. The analysis, by the Intergovernmental Panel on Climate Change (IPCC) [2], has indicated that the capture and storage of  $CO_2$  would contribute between 15 and 55% to the global cumulative mitigation effort until 2100, thus presenting itself as a transition technology that will contribute to mitigating climate change.

Various methods for reducing the amount of  $CO_2$  in the atmosphere have been discussed, such as reducing the amount of energy used and improved efficiency of energy generating processes, use of non-fossil fuels (renewable energy),  $CO_2$  capture, and sequestration.

The capture of  $CO_2$  is part of the set of technologies that try to separate  $CO_2$  from large emission sources (thermal power plants, cement plants, steel mills, ceramics, etc.) and prevent their emission into

the atmosphere. The first solutions for the capture of  $CO_2$  implied the need to inseparably link the catch with transport and geological storage (capture, transport and geological storage technologies (CCS) [3]), the latest research on these technologies is directed towards the generation of a transportation network and geological storage of  $CO_2$ . This network would unite different geographical points of geological storage with enough capacity and would allow decoupling the capture of  $CO_2$  from transport and geological storage.

Bui et al. (2018) [4] have reviewed the current status of CO<sub>2</sub> capture, transport, utilization, and storage from a multi-scale perspective. Deployment of large-scale CCS projects has been slow. Reiner 2016 [5] indicated that the implementation of CCS technologies worldwide will require a portfolio of large-scale demonstration projects. Of the 37 major large-scale CCS projects, 17 of these are in operation, four in construction and the remainder are in varying stages of development [4,6]. In addition, Bui et al. (2018) [4] have contextualized CCS technologies in integrated evaluation models (IAM). In order to limit global warming, negative emission (NET) technologies consisting of combining CCS with low-carbon or carbon-free Bioenergy (BECCS) [4,7] are considered key. BECCS has the double benefit of mitigating emissions and generating energy, which makes it attractive from the perspective of cost optimization of an IAM.

Moreover, the knowledge of thermodynamic and transport properties of CO<sub>2</sub>-mixtures is important for designing and operating different processes in carbon capture and storage systems (Tan et al. (2016) [8]).

Other capture technologies consist of giving use to captured  $CO_2$ , these are the so-called carbon capture and utilization (CCU) technologies [9]. One of the advantages of CCU over CCS is that utilization of  $CO_2$  is normally a profitable activity as products can be sold. In addition, storage potential may be limited or away from  $CO_2$  sources and public resistance to geological storage of  $CO_2$  has been noted [10]. Baena-Moreno et al. (2015) [11] presented a list of technologies for Carbon Capture and Utilization (CCU) from laboratory scale to commercially established  $CO_2$  applications, such as raw materials for the manufacture of fuels, polymers, Chemicals, or as recovery agents in techniques such as Enhanced Oil Recovery (EOR) and Enhanced Coal Bed Methane (ECBM). These technologies not only seek to reduce the volume of  $CO_2$  emissions into the atmosphere, but also obtain a benefit using  $CO_2$  in different types of industrial processes replacing conventional raw materials and reducing the consumption of fossil fuels.

Another form of CCU would be biological carbon mitigation (BCM), which consists of the capture of  $CO_2$  by photosynthetic organisms to convert it into organic carbon using solar energy and producing a significant amount of biomass. BCM is a way to sequester  $CO_2$  while producing biofuels and valuable by-products. Microalgae serve as a promising agent for carbon sequestration [12,13].

Among the capture systems proposed by the report of the Intergovernmental Panel on Climate Change (IPCC) published in 2005 [3], the most mature and widely used post-combustion capture technology is chemical absorption with monoethanolamine (MEA), which however is an important contributor to CO<sub>2</sub> emissions and global warming. Therefore, one of the proposed solutions would be the search for more environmentally sustainable adsorbents (Cuellar et al. (2015) [9]). Bui et al. (2018) [4] highlight three aspects of the post-combustion capture of the calcination/carbonation cycles with calcite compared to other technologies: one, that the carbonator / calciner can serve as a heat source for a steam cycle to produce energy additional, another, that the absorber used, calcite, is available in industrial quantities and is a non-hazardous chemical, and finally, that there is the possibility of using spent absorbent in industrial processes such as cement manufacturing, with the consequent decarbonisation partial of this industry or can even achieve near zero emissions by incorporating this technology into its manufacturing process. Calcium looping (CaL) technology has also been progressed to pilot scale. There are two major demonstration projects, one at the University of Darmstadt, Germany [14], and one in La Pereda, Spain [15].

 $CO_2$  is normally accompanied, in the gas stream produced from the combustion of coal, by  $SO_2$ ,  $NO_x$ , and  $O_2$ . The use of the carbonation reaction of CaO to CaCO<sub>3</sub> as a possible separation reversible

process of CO<sub>2</sub>, is adequate, although CaO will also react with SO<sub>2</sub> to form CaSO<sub>4</sub> [3]. Calcination, to regenerate the adsorbent, takes place at high temperatures, at around 900–950 °C. In carbonation, a flow of CO<sub>2</sub> gas is in contact with CaO, at temperatures above 600 °C, with the carbonation reaction being carried out to form CaCO<sub>3</sub>. [16]. The maximum carbonation capacity of CaO decays with the number of carbonation/calcination cycles down to 20% [17,18]; it is necessary to indicate that these authors used particle sizes greater than 100  $\mu$ m.

Reactions of CaO with SO<sub>2</sub> could be a problem in the calcination/carbonation cycles since CaSO<sub>4</sub> does not decompose at the calcination temperature [19]. This implies an accumulation of CaSO<sub>4</sub> in the cycles, which reduces the efficiency of carbonation [20]. The separation of CaSO<sub>4</sub> is only possible through a purge with a consequent loss of sorbent [19]. Consequently, SO<sub>2</sub> should be removed from the gas stream prior to the incorporation of the gas stream into the calcination/carbonation cycles. This is now usual in coal-fired power plants, since by regulation (DEI, 2010) [21], they have to incorporate a desulphurization system. Consequently, if the SO<sub>2</sub> concentration is very low, the residence time is short, and the temperature remains below 640 °C [22], the carbonation reaction will take place with the sulphation reaction being unlikely.

Different studies about reversibility show the carbonation is far from being reversible [11,12]. After a very fast, and chemically controlled initial reaction, there is a much slower second phase controlled by diffusion into  $CaCO_3$  layers [23–25]. Several papers [20,23,26–28] show a decrease in the carbonation maximum of lime with the number of cycles and this has been related to sintering within the material.

Several authors [29–32] have shown that the formation of a dense, non-porous layer of carbonate minerals takes place around the reagent particles, so that total carbonation efficiency cannot be attained, except for cases of high relative humidity and small particle sizes (nanometric scale). This protective layer causes a decrease in porosity when the materials are partially carbonated [31–33]. This is in agreement with the poor results obtained by Escosa et al. [19] regarding the reaction between CaO and  $CO_2$  that requires the introduction of much more CaO than is necessary stoichiometrically.

Carbonation can be improved under the action of water at moderate temperature and low  $CO_2$  pressure [34]. The hydration of CaO with water vapor is typically used to reactivate its performance as a sorbent [35]. In this sense, Maretic, et al. [36], indicated that carbonation was much greater when Ca(OH)<sub>2</sub> is directly carbonated at high temperatures compared to carbonation of CaO. The carbonation reaction of nanoparticles [35] that takes place is:

$$Ca(OH)_{2(s)} + CO_{2(g)} = CaCO_{3(s)} + H_2O_{(g)}$$
(1)

The molecular water produced in the reaction is not stable at the hydroxide-carbonate interface [35]. This assumption is only valid for small Ca(OH)<sub>2</sub> particles (<30  $\mu$ m) [37]. Consequently, the results show that the simultaneous expulsion of the molecular water, produced during the carbonation of Ca(OH)<sub>2</sub>, could significantly improve the transfer of CO<sub>2</sub> from the gas phase to the surface of the unreacted Ca(OH)<sub>2</sub>. If CaO is reacted with CO<sub>2</sub> at the very moment of the decomposition of Ca(OH)<sub>2</sub>, with particle size less than 30  $\mu$ m, the adsorption is at a maximum [35].

In order to develop good  $CO_2$  adsorbents at high temperature several aspects need to be investigated: understanding the mechanism of  $CO_2$  adsorption on different kinds of adsorbents, finding the best adsorbent for modifying chemical reagent to improve the adsorption capacity; study of the adsorption and desorption kinetics of  $CO_2$  in the adsorbent and the choice of optimal conditions for the industry [38].

The present work deals exclusively with the capture phase, leaving open any of the following possibilities of action with the captured  $CO_2$ , from its transportation and geological storage, to which it no longer has to be linked in an inseparable way, since it can be part of a transport network towards different possibilities of confinement or it can be used with good performance in any of the possible alternatives described by Baena-Moreno et al. (2015) [11].

This paper is intended to address  $CO_2$  capture using natural limestone (CaCO<sub>3</sub>) of different granulometries, calcined limestone (lime (CaO)) and calcined and rehydrated limestone (portlandite (Ca(OH)<sub>2</sub>)). Among other aspects, the particle size variation in several adsorption-desorption cycles and how this affects the adsorption-desorption temperature and the amount of  $CO_2$  adsorbed is discussed. Finally, the stability of the adsorbent samples after a large number of adsorption-calcination cycles is considered.

# 2. Materials and Methods

Nine samples were analysed. The first three A15, A1045, A1M1, are from 0–6000  $\mu$ m size fraction of crushed limestone from the quarry "Horcallana" owned by Endesa (Teruel, Spain) (sample A1). The sample A15 corresponds to less than 500  $\mu$ m size fraction of sample A1. A1045 sample corresponds to the particle size less than 45  $\mu$ m from the sample A1. The A1M1 sample corresponds to the size fraction below the 100  $\mu$ m A1 sample. The samples were subjected to grinding in a laboratory mill with porcelain balls (45 balls of 30 mm in diameter and each weighing 28 g) for 4 h.

Three uncalcined samples were calcined at 900 °C for 1 h. Samples were identified by adding a "C" to the name of the uncalcined sample (A15C, A1045C, A1M1C). Then, three samples were partially rehydrated in an atmosphere rich in water. Samples were identified by adding an "H" to the name of the calcined sample (A15CH, A1045CH, and A1M1CH). In Scheme 1, the nomenclature and processes of the samples are summarized.



Scheme 1. Nomenclature and processes of the samples.

PXRD patters were recorded in a Siemens D-5000 instrument (Billerica, MA, USA), using CuK $\alpha$  radiation ( $\lambda = 1.5418$  Å) working at a voltage of 40 kV and a current 30 mA. The detector is a scintillation counter with a slit of 1 mm, employing a scan rate of 2° min<sup>-1</sup> using a step of 0.2° every 0.6 s and sweeping angles between 10 and 80° in units of 2 $\theta$ .

Particle size data were obtained in a particle size analyser by laser diffraction technology 2000F Mastersizer (Malvern, UK) with one dry dispersion unit Venturi Scirocco 2000 of Malvern Instruments (Malvern, UK).

The thermogravimetric analysis was carried out in a Setaram Setsys Evolution 16/18 apparatus (Caluire, France), at a heating rate of 5  $^{\circ}$ C min<sup>-1</sup> in air.

 $N_2$  isotherms were determined on an Autosorb iQ (Quantachrome) (Boynton Beach, FL, USA), with samples being degassed at 150 °C under vacuum for 2 h prior to this. The surface was calculated by applying the BET (Brunauer, Emmett and Teller) method in the range of relative equilibrium pressure  $0.05 \le P/P_0 \le 0.20$  [39]. The porosity distribution was determined by a density functional theory (DFT) method using a cylindrical model for  $N_2$  on an oxide surface.

Microstructural characterisation of the materials was carried out using JEOL 1400 TEM (Jeol, Tokyo, Japan).

The CO<sub>2</sub> adsorption studies were carried out in a Setaram Setsys Evolution 16/18 apparatus (Caluire, France). 40.00 mg of sample was introduced into the equipment and it was evacuated before a CO<sub>2</sub> flow of 20 mL min<sup>-1</sup> was passed for one hour, with the thermogravimetric analysis being carried out up to a temperature of 1000 °C.

Finally, CO<sub>2</sub> adsorption isotherms were recorded on a Micromeritics ASAP 2020 apparatus (Norcross, GA, USA). Commonly used to obtain the BET isotherms of N<sub>2</sub>, which has the possibility of using CO<sub>2</sub> adsorbate gas in its software, using samples previously outgassed at 150 °C under vacuum for 2 h. The CO<sub>2</sub> adsorption isotherms were carried out at 0 °C between the values of 0–0.035 of P/P<sub>0</sub>.

The study of CO<sub>2</sub> adsorption in the different samples was carried out by means of four different procedures (Scheme 2). In the first, the CO<sub>2</sub> adsorption isotherms were performed in a Micromeritics ASAP 2020 apparatus, commonly used to obtain the BET isotherms of N<sub>2</sub>, which has the possibility of using CO<sub>2</sub> adsorbate gas in its software. In the second, the samples were subjected to a CO<sub>2</sub> flow of 20 mL min<sup>-1</sup> for 1 h in a thermogravimetric analysis equipment (ATD-TG) (Caluire, France) and later a thermogravimetric analysis (ATD-TG) was carried out in an air atmosphere up to 1000 °C. In the third, successive adsorptions of CO<sub>2</sub> were carried out to determine the fall of the adsorption capacity after a certain number of cycles. For this procedure the sample, calcined in a previous step, was placed inside in the TG/DTA apparatus and then the sample was treated with a CO<sub>2</sub> flow of 20 mL min<sup>-1</sup> for one hour and finally, thermal analysis up to 1000 °C was conducted. The fourth procedure consisted of the simultaneous heating of two samples, one calcined and one totally hydrated, in a tubular oven in a CO<sub>2</sub> atmosphere produced by a CO<sub>2</sub> flow of 20 mL min<sup>-1</sup> during the whole calcination process (1/2 h) at a temperature of 550 °C.



Scheme 2. Procedures for the study of CO<sub>2</sub> adsorption in different samples.

#### 3. Results and Discussion

#### 3.1. Characterization of Samples

The X-ray diffraction diagrams of the original samples without calcining (Figure 1A), shows the presence of calcite (JCPDS: 05-0586) [40] as a main phase, with a negligible amount of impurities of SiO<sub>2</sub> (JCPDS: 46–1045) [40]. In the samples calcined at 900 °C for 1 h (Figure 1B), CaO was the main phase detected (JCPDS: 37–1497) [40] with negligible amounts of SiO<sub>2</sub> (JCPDS: 46–1045) [40] and Ca<sub>2</sub>SiO<sub>4</sub> (JCPDS: 33-0303) [40].



**Figure 1.** X-ray diffraction (XRD) patterns. (**A**) calcites: A1045, A1M1, and A15; (**B**) calcites calcined: A1045C, A1M1C, and A15C.

Figure 2 shows the particle size distribution curves of the original and calcined samples. In the sample A15, it can be seen that the calcination leads to a considerable decrease in the particle size, with the maximum of the particle size distribution decreasing from around 158 to around 89  $\mu$ m. The calcination had a minimal effect on the sample size A1045 (maximum at around 22  $\mu$ m). The sample A1M1, subjected to double-milling, has a bimodal distribution of the particle size, with its average size being the lowest of all (19  $\mu$ m). Calcining gives rise to an increase in the proportion of smaller particle size, although the average distribution shows little change, around 18  $\mu$ m. In this case, as well, the distribution peak of larger particle size is around 28  $\mu$ m.



Figure 2. Grain-size distribution of calcites: A1045, A1M1, A15 and calcites calcined: A1045C, A1M1C, and A15C.

Figure 3 (left) shows the  $N_2$  adsorption isotherms of the three original samples (CaCO<sub>3</sub>), calcined (CaO) and rehydrated to Ca(OH)<sub>2</sub>. Their isotherms were of type IV with a sharp increase in adsorption at  $P/P_0$  between 0.05 and 0.45 due to capillary condensation, typical of materials with a narrow distribution in the mesopore range (Figure 3 (right)). For the three calcined and rehydrated samples, a sharp capillary condensation step at  $P/P_0 = 0.45$  was observed. All samples present a hysteresis loop of type H3 or H4, which does not exhibit any limiting adsorption at high  $P/P_0$ , related to aggregates of plate-like particles. The three calcined and rehydrated samples present a hysteresis loop of type H3 with slit-shaped pores. The rest of the samples present a hysteresis loop of type H4 often associated with narrow slit-like pores. This is in accordance with the pore size distributions (Figure 3 (right)). The main physical properties of the three adsorbents are summarized in Table 1. The specific surfaces obtained were 2.7, 4.7, and 3.1 m<sup>2</sup> g<sup>-1</sup> for the samples A15, A1045 and A1M1 respectively, and 4.4, 4.5 and 3.9  $m^2$  g<sup>-1</sup> for the calcined samples A15C, A1045C, and A1M1C, respectively. While the surface area of the three calcined samples is very similar, the largest initial average size increase was for the A15 sample, the one that has the largest initial size. A very significant increase is observed in the calcined and rehydrated samples A15CH, A1045CH and A1M1CH, of 21.3, 13.8 and 18.9 m<sup>2</sup>  $g^{-1}$ , respectively. This results in a significant increase in pore volume of  $8.84 \times 10^{-2}$ ,  $5.70 \times 10^{-2}$  and  $5.87 \times 10^{-2}$  cm<sup>3</sup> g<sup>-1</sup>, respectively (Table 1 and Figure 3 (left)). The average pore diameter was calculated between 2.77–3.97 nm (Table 1 and Figure 3 (right)). It is important to note that in the two calcinated and hydrated samples of smaller size, the same average pore diameter value (3.97 nm) is observed. A negligible micropore content can also be observed for the calcined samples and for the calcined and rehydrated ones (Figure 3 (right)).

Table 1. Physicochemical properties of pristine calcites, calcined and rehydrated.

Material	$S_{BET}^{a} (m^2 g^{-1})$	$S_{mp}^{\ \ b} (m^2 g^{-1})$	$V_p \ ^c (10^{-2} \ cm^3 \ g^{-1})$	D <sub>p</sub> <sup>d</sup> (nm)	$S_{eq} e (m^2 g^{-1})$
A15	2.7	0	0.82	2.89	10.8
A15C	4.4	0	0.99	2.89	12.9
A15CH	21.3	0	8.84	3.79	18.4
A1045	4.7	0	1.72	2.77	10.9
A1045C	4.5	0	1.16	2.77	11.9
A1045CH	13.8	0	5.70	3.97	15.4
A1M1	3.1	0	0.88	2.89	10.4
A1M1C	3.9	0	0.92	3.17	12.1
A1M1CH	18.9	0	5.87	3.97	19.1

<sup>a</sup> BET specific surface area determined in the range of relative pressures from 0.05 to 0.2; <sup>b</sup> Micropore area, calculated by t-Plot method. <sup>c</sup> Single-point pore volume; <sup>d</sup> Diameter pore average size, calculated from the adsorption branch according to the BJH method; <sup>e</sup> Calculated from  $CO_2$  isotherm.

Volume STP (cm<sup>3</sup>/g)



0,2 0,4 0,6 0,8 1 0 Relative Pressure (P/Po) Half pore width (Å) Figure 3. (Left) Nitrogen adsorption-desorption isotherms for calcites: A1045, A1M1, and A15; calcites

5 10 15 20 25 30 35 40 45 50

calcined: A1045C, A1M1C, A15C and calcites calcined and rehydrated: A1045CH, A1M1CH, and A15CH. (Right) Pore size distributions for calcites: A1045, A1M1, and A15; calcites calcined: A1045C, A1M1C, A15C, and calcites calcined and rehydrated: A1045CH, A1M1CH, and A15CH.

#### 3.2. CO<sub>2</sub> Adsorption Studies

In order to delve further into the possibilities of adsorption of original, calcined or hydrated calcites, in line with its possible use for capturing  $CO_2$  in a power plant, four tests were carried out that allow us to elucidate which is the optimum particle size, as well as the conditions of adsorption. In the first test, the CO<sub>2</sub> adsorption isotherms were conducted at 0  $^{\circ}$ C between the values of 0–0.035 of P/P<sub>0</sub>. In all the calcined samples (A15C, A1045C, and A1M1C) an increase of the equivalent surface area is observed, with respect to the samples without calcining (A15, A1045, and A1M1), and especially in the partially calcined and rehydrated samples (A15CH, A1045CH and A1M1CH) (Table 1). These results show a greater capacity of CO<sub>2</sub> adsorption of the A1M1CH sample (5.3 mg of CO<sub>2</sub>  $g^{-1}$  of adsorbent) (Figure 4), which is in accordance with the highest value of equivalent surface  $(19.1 \text{ m}^2 \text{ g}^{-1})$  (Table 1). The results obtained show that the capture of  $CO_2$  for the sample with a smaller average particle size, A1M1, is more effective. To complete these results, an additional experiment was carried out at room temperature and high CO<sub>2</sub> pressure, which introduced 50.00 gr of freshly calcined A1M1 sample (at 900 °C for 2 h in an air atmosphere) in a three-litre volume, metal bottle, for 24 h at room temperature and at a pressure of 23 bars of  $CO_2$ . Subsequently, the solid was extracted and an increase of 0.375 g (0.75%) was noted. In short, 7.9 mg of  $CO_2 g^{-1}$  of adsorbent was adsorbed, which is consistent with the results obtained by other authors, e.g., Yong et al. [38]. XRD diagrams (not shown) confirmed that the only phase present before and after the experience was CaO, which it is in accordance with the physical nature of the adsorption.



**Figure 4.** CO<sub>2</sub> adsorption–desorption isotherms for calcites: A1045, A1M1, and A15; calcites calcined: A1045C, A1M1C, A15C, and calcites calcined and rehydrated: A1045CH, A1M1CH, and A15CH.

In tests 2 and 3 we analyzed the effect of high temperature and the effect of adsorption-calcination cycles on the capacity of  $CO_2$  adsorption.

The results of the second test are shown in Figure 5 and in Table 2. In calcite samples that were not subjected to a flow of  $CO_2$ , an endothermic effect is observed, which is associated with the transformation of calcite (CaCO<sub>3</sub>) in lime (CaO), centred around 840–857 °C. A slight decrease in the decomposition temperature is observed as the average size of the particles decreases (Figure 2). In

all the samples subjected to a current of  $CO_2$ , a stabilization of the calcite is observed since it slightly increases the temperature of the endothermic effect, 874–907 °C., being larger in samples of larger particle size. This is related to the equilibrium between the  $CO_2$  in the atmosphere surrounding the particulates and the mineralized  $CO_2$  on the surface thereof. The reaction will only proceed if the partial pressure of  $CO_2$ , in the gas above the solid surface, is less than the decomposition pressure of the CaCO<sub>3</sub> [41,42].



**Figure 5.** TG (solid lines) and DTA (dotted line) curves: for calcite A15, calcite A15 +  $CO_2$ , calcite calcined A15C +  $CO_2$  and calcite calcined and rehydrated A15CH +  $CO_2$ ; for calcite A1045, calcite A1045 +  $CO_2$ , calcite calcined A1045C +  $CO_2$  and calcite calcined and rehydrated A1045CH +  $CO_2$ ; for calcite A1M1, calcite A1M1 +  $CO_2$ , calcite calcined A1M1CH +  $CO_2$  and calcite calcined and rehydrated A1M1CH +  $CO_2$ .

**Table 2.** Gas-solid carbonatation of different samples under non-isothermal conditions at the same heating rate (5 °C min<sup>-1</sup>).

Sample	∆weight (%)	∆weight <sup>a</sup> (%)	Total Loss of Water (%)	Total loss (%)	Ca(OH) <sub>2</sub> (%)	CaO <sup>b</sup> (%)	CaO <sup>c</sup> (%)
A15	-41.81	-44.00	-	-41.81	-		-
A15C	+26.66	+78.48	-	-26.66	-	100.00	33.97
A15CH	+26.07	+64.68	-7.72	-33.79	31.80	68.20	27.85
A1045	-41.12	-44.00	-	-41.12	-		-
A1045C	+32.98	+78.48	-	-32.98	-	100.00	33.51
A1045CH	+25.64	+66.88	-6.50	-32.14	26.75	73.25	28.28
A1M1	-41.86	-44.00	-	-41.86	-		-
A1M1C	+33.05	+78.48	-	-33.05	-	100.00	42.11
A1M1CH	+28.39	+63.39	-8.45	-36.84	34.80	65.20	31.61

<sup>a</sup> Theoretical value; <sup>b</sup> CaO total; <sup>c</sup> CaO reacted from total of CaO.

For the calcined samples (A15C, A1045C, and A1M1C) exposed to a current of  $CO_2$ , a continuous weight gain in a single stage is observed which correlates with the formation of CaCO<sub>3</sub>. In the calcined and partially rehydrated samples (A15CH, A1045CH, and A1M1CH) a weight increases in two stages, related to the formation of CaCO<sub>3</sub> is seen. The final mass is lower than the starting mass and the difference in weight correlates with the water loss from Ca(OH)<sub>2</sub> in its transformation to CaO.

The results obtained for the calcined samples (A15C, A1045C and A1M1C) are in agreement with those obtained for non-isothermal conditions by Montes-Hernandez et al. [35], although lower than those obtained for these authors. This is because the experimental work conditions used were different, Montes-Hernandez, et al. [35] maintained a  $CO_2$  flow of 50 mL min<sup>-1</sup> throughout the experiment. Total carbonation is achieved at temperatures above 850 °C. In the present case, the greatest increase in weight, due to Equation (2), was somewhat less than half the theoretical maximum amount according to Equation (2). In other words, 42.11% of the total CaO was carbonated. The total adsorption of  $CO_2$  was greater in the A1M1C sample and corresponds to a value of 330.5 mg of  $CO_2$  g<sup>-1</sup> of adsorbent.

$$CaO_{(s)} + CO_{2(g)} \rightarrow CaCO_{3(s)}$$
<sup>(2)</sup>

Regarding the calcined and partially rehydrated samples (A15CH, A1045CH, and A1M1CH), the difference between the total weight increase and the total weight loss enables the calculation of the amount of water corresponding to  $Ca(OH)_2$  and therefore, the amount of  $Ca(OH)_2$  and CaO present in the samples. The increase in weight in the first stage corresponds to Equation (1), with the largest increase occurring in the sample of smaller average particle size A1M1CH, 17.57%. In the A15CH and A1045CH samples, the increase was 14.56 and 14.59, respectively.

This reaction assumes the simultaneous release of molecular water from Ca(OH)<sub>2</sub>, and the incorporation of CO<sub>2</sub> into the particles with the subsequent carbonation, mainly for small particles of Ca(OH)<sub>2</sub> (<30  $\mu$ m), Blamey et al. [37]. In the present work, this process occurs even for sizes up to 89  $\mu$ m (sample A15C), although the adsorption of CO<sub>2</sub> is somewhat lower. This behavior seems to be in accordance with that established by Marsh and Ulrichson [22] that indicates that the hydration of CaO to Ca(OH)<sub>2</sub> and subsequent calcination to CaO, gives rise to a much more porous adsorbent.

The second weight increase corresponds to the carbonation reaction of CaO (Equation (2)). The total weight increase was greater in the A1M1CH sample (28.39%). Subsequently, the decomposition of the formed calcite (CaCO<sub>3</sub>), occurs at a temperature slightly lower than that of the original samples. These results are in agreement with the studies carried out by Montes et al. [35] that showed that the carbonation of CaO occurs in a wide range of temperatures (100–900 °C) and that of Ca(OH)<sub>2</sub> occurs in a single narrow range of temperatures (500–600 °C). The adsorption was lower than those obtained by Montes-Hernandez et al. [35]. This is because the experimental work conditions used by these authors were different, maintaining a CO<sub>2</sub> flow of 50 mL min<sup>-1</sup> throughout the experiment. The total adsorption of CO<sub>2</sub> was greater in the A1M1CH sample and corresponds to a value of 368.44 mg of CO<sub>2</sub> g<sup>-1</sup> of adsorbent showing that approximately, 32% of the total CaO has reacted. The presence of CaO partially hydrated (A1M1CH) favors the adsorption of CO<sub>2</sub>, compared to the unhydrated sample (A1M1C) for which a value of 330.5 mg of CO<sub>2</sub> g<sup>-1</sup> of adsorbent was obtained. Further, that is in accordance with the surface increase (S<sub>BET</sub>) and of pore volume (V<sub>p</sub>) (Figure 1) that occurs when hydrating of the sample A1M1C. That concurs with the results commented by Marsh and Ulrichson [22].

The mechanism of carbonation has been proposed by different authors (Barker [23], Bhatia [24], Mess [25], and Abanades [16]). After a very rapid, chemically controlled, initial reaction period a much slower second stage is followed, controlled by diffusion in the CaCO<sub>3</sub> layers. It is for this reason that the smaller the particles, the faster the diffusion stage and the deeper the carbonation, reaching, in some cases, the total carbonation of the particles.

The third test to be analyzed was the loss of adsorption capacity after several cycles adsorption-calcination in an ATD-ATG apparatus. The samples with larger and smaller particle sizes, A15 and A1M1, were chosen for analysis to see if particle size would play an important role in the adsorption process. Figure 6 include the ATD-TG diagrams of the A15 and A1M1 samples after 5 and 30 cycles. After the calcination process of each cycle, the only phase present was CaO (X-ray diffraction diagrams not shown). According to the ATG curves, there is a considerable reduction in the temperature at which the elimination of CO<sub>2</sub> takes place, from 857 °C to 644 °C (Figure 6 left) and from 840 °C to 682 °C (Figure 6 right), compared to the original samples after 5 cycles, and to 644 °C (Figure 6 left) and 702 °C (Figure 6 right) after 30 cycles. Likewise, the results show that repeated calcination

at 900 °C, without previous hydration steps, reduced the capacity of CO<sub>2</sub> adsorption considerably. This effect is less pronounced (Figure 6 right) when the particle size is smaller (Figure 2). In this case, the A1M1C sample, after 30 cycles, retains an adsorption capacity of 3.83% (30 mg of CO<sub>2</sub>  $g^{-1}$  of adsorbent), higher than that of sample A15C, which is only 0.55% (4.3 mg of CO<sub>2</sub>  $g^{-1}$  of adsorbent). This agrees with the transmission microphotographs (Figure 7) of A15C and A1M1C samples after 5 and 30 cycles. In all cases, homogeneous layered particles with hexagonal habits were observed, with low porosity and slightly larger grain size for A15C sample. These results are in agreement with the studies carried out by Abanades [16], Abanades and Alvarez [17], Grasa, et al. [18], and Blamey et al. [43] who observed that the initial capacity of limestone adsorption/desorption after multiple cycles fell to an asymptotic value as a consequence of sintering processes. Furthermore, the lower the temperature is, and the less time spent in the calciner, the less the sintering occurs. These results seem to indicate that, after the first adsorption-calcination cycle, the calcination temperature decreased to around 700 °C, which reduces the risk of sintering.



**Figure 6.** TG (solid lines) and DTA (dotted line): (left) curves for calcite A15, calcite A15 after 5 cycles and calcite A15 after 30 cycles of calcination— $CO_2$  flow of 20 mL min<sup>-1</sup> (1 h)—calcination in air a 900 °C; (right) curves for calcite A1M1, calcite A1M1 after 5 cycles and calcite A1M1 after 30 cycles of calcination— $CO_2$  flow of 20 mL min<sup>-1</sup> (1 h)—calcination in air a 900 °C;



**Figure 7.** TEM images for calcites A15 and A1M1 after 5 and 30 cycles of calcination—CO<sub>2</sub> flow of 20 mL min<sup>-1</sup> (1 h)—calcination in air a 900 °C.

The fourth test consisted of a study of CO<sub>2</sub> adsorption in a tubular oven using the calcined A1M1 sample (A1M1C) and calcinated and rehydrated (A1M1CH). The diffractograms (Figure 8) show the presence of CaO in the first and Ca(OH)<sub>2</sub> (JCPDS: 04-0733) [28] in the second. Once placed in the tubular oven (10 g), a CO<sub>2</sub> atmosphere was established for which a flow of CO<sub>2</sub> was passed through the tubular furnace during the whole experiment (1/2 h), and the temperature was raised to 550 °C (non-isothermal conditions), trying to simulate what would happen in an industrial process if the adsorbent is put in contact with a stream of CO<sub>2</sub>. After 30 min, as reflected in the X-ray diffraction patterns (Figure 8), only in the calcined and rehydrated sample was the carbonation process total, since all the X-ray reflections correspond to the calcite phase (CaCO<sub>3</sub>). The carbonation of CO<sub>2</sub> in the newly created CaO surface (Equation (1)).



**Figure 8.** X-ray diffraction (XRD) patterns of the samples A1M1C and A1M1CH and the samples A1M1C and A1M1AC heated at 550 °C in a tubular oven with a CO<sub>2</sub> flow of 20 mL min<sup>-1</sup>.

In the experimental conditions used, the activation energy is not sufficient to cause the total carbonation of CaO, since higher temperatures are needed [35], and instead are only sufficient to carbonate the Ca(OH)<sub>2</sub> at the same moment it decomposes to CaO (481 °C). This is in agreement with the studies of Materic, et al. [36] who indicated that solid-gas carbonation was much greater when Ca(OH)<sub>2</sub> was directly carbonated at intermediate temperatures, compared to carbonation of CaO. On the other hand, when the reaction occurs at lower temperatures, the risk of sintering is very low [43], with the subsequent recyclability of the adsorbent for further cycles.

The weight increases for sample A1M1CH, treated in a tubular oven, was 34.11%, very close to the theoretical capacity of 35.11% and the adsorption capacity was 577.00 mg  $g^{-1}$  close to the theoretical

maximum of 593.95 mg g<sup>-1</sup>. This result is in agreement with those reported by Montes-Hernández, et al. [35], proposing a temperature greater than 300 °C for the carbonation reaction. These authors point out that at temperatures lower than 600 °C under isothermal conditions they never reach total carbonation levels using Ca(OH)<sub>2</sub> nanoparticles and attribute it to the formation of dense, non-porous sheets of mineral carbonate around the particles of Ca(OH)<sub>2</sub>. Under non-isothermal conditions they achieve total carbonation at temperatures close to 600 °C, as already mentioned in relation to the simultaneous exit of H<sub>2</sub>O and carbonate formation (Blamey et al. [37]).

The experiment was repeated 10 times with the same sample A1M1, calcined and rehydrated. The calcite phase (CaCO<sub>3</sub>) is the only phase present after each carbonation cycle. No appreciable losses of adsorption capacity were observed. This is in agreement with Nikulshina et al. [44] who reported that after five cycles of carbonation-calcination, the loss of adsorption capacity was 0.1%. The carbonation was carried out at 365–400 °C and the calcination at 800–875 °C and virtually no sintering occurs.

In summary, it can be concluded that limestone samples of average particle size  $<30 \mu m$  are the ones that have the most favourable size for CO<sub>2</sub> capture. The previous hydration of the oxides formed, and the temperature of 550 °C, favours the CO<sub>2</sub> capture process.

#### 4. Conclusions

This work has analyzed the influence of particle size of a calcite from a quarry on the efficiency of CO<sub>2</sub> capture. The final objective is its application in the elimination of CO<sub>2</sub> from gases emitted by coal-fired power plants. Three different size fractions of particle  $<500 \mu$ m,  $<100 \mu$ m, and  $<45 \mu$ m have been studied. Calcite, calcined calcite and calcined and rehydrated calcite were used and were characterised by means of X-ray diffraction, N<sub>2</sub> adsorption isotherms, thermogravimetric analysis, transmission electron microscopy and particle size by laser diffraction. The calcination process had little effect on the particle size of the smaller samples A1045 and A1M1 ( $<30 \mu$ m). The N<sub>2</sub> and CO<sub>2</sub> isotherms showed a very important increase in the surface of the calcined and rehydrated samples (A15CH, A1045CH, and A1M1CH) with respect to the calcinated or pristine samples. Likewise, the same value of average pore diameter (3.97 nm) was observed in the samples A045CH and A1M1CH.

The results of the  $CO_2$  isotherms show that the capacity of physical adsorption of  $CO_2$  using calcite, calcined calcite or calcined and rehydrated calcite is very low, although the particle size was small (<30  $\mu$ m). The best results were obtained for the sample A1M1CH (5.3 mg of CO<sub>2</sub> g<sup>-1</sup> of adsorbent). Even when the pressure is increased to 23 bars of CO<sub>2</sub>, the capture is low, the sample A1M1C adsorbed 7.9 mg of  $CO_2$  g<sup>-1</sup> of adsorbent. This same sample at the pressure of 1 bar only captured 2.3 mg of  $CO_2$  $g^{-1}$  of adsorbent. By contrast, chemical capture at high temperature was more efficient. Regarding calcined samples exposed to a current of  $CO_2$ , the largest adsorption of  $CO_2$  was for the sample with the smallest particle size A1M1C (330.5 mg of  $CO_2$  g<sup>-1</sup> of adsorbent). In calcined and partially rehydrated calcites the adsorption improves, the total adsorption of  $CO_2$  was greater in the sample A1M1CH (368.44 mg of  $CO_2 g^{-1}$  of adsorbent). The repeated calcination at 900 °C, without previous hydration steps, reduces the capacity of  $CO_2$  adsorption considerably. This effect was less pronounced when the particle size was smaller, where the sample A1M1C, after 30 cycles, retained an adsorption capacity of 30 mg of CO<sub>2</sub>  $g^{-1}$  of adsorbent, higher than that of sample A15C, which was only 4.3 mg of CO<sub>2</sub>  $g^{-1}$  of adsorbent. Finally, when the adsorption of  $CO_2$  was carried out in a tubular furnace at 550 °C, just at the moment when the  $Ca(OH)_2$  was decomposed, the adsorption was practically total (577 mg of  $CO_2$  g<sup>-1</sup> of adsorbent). When the calcined sample is completely rehydrated after each cycle, the loss of adsorption capacity is practically negligible, thus improving the recyclability of the adsorbent for subsequent cycles.

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