

Article

Estimating the Annual Above-Ground Biomass Production of Various Species on Sites in Sweden on the Basis of Individual Climate and Productivity Values

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Abstract: The literature contains a large number of bioclimate, climate and biometric models for estimating the production of different species or stands under specific conditions on a defined site or models giving the distribution of a single species. Depending on the model used, the amount of input data required varies considerably and often involves a large investment in time and money. The purpose of this study was to create a model to estimate the annual above-ground biomass production of various species from site conditions defined by mean annual temperature and mean annual precipitation. For this approach, the Miami model of Lieth was used as a base model with some modifications. This first version of the modified model was restricted to sites in Sweden, where changes in the soil and groundwater level were relatively small, and where the growth of land vegetation was mostly dependent on temperature. A validation of this model has shown that it seems possible to use the Miami model to estimate the annual above-ground biomass production of various species, and that it was possible to compare the annual above-ground biomass production of different species on one site, as well as the annual above-ground biomass production of different species on different sites using the modeled data.

Keywords: MAT; MAP; bioclimatic model; above-ground biomass production

1. Introduction

The growth and geographical distribution of species are influenced by both abiotic and biotic factors. Abiotic factors for plants, also known as physical and chemical factors, are light, climate and geology [1,2]. Biotic factors are the effects of living organisms on the environment, such as the competition for resources and space, or natural enemies [3]. It is clear that no single factor affects the survival, development and fitness of plants, but that there is an optimum for each factor where plants grow best [4]. An increased distance from the optimum leads to greater stress, until the limit of tolerance is reached and death occurs. As there is no single optimum, but rather a number of different optima for different factors such as temperature, light, moisture, nutrients, *etc.*, and as they are all available to different extents, there is always at least one factor that limits the growth; the one that is least available. This is in essence “the law of limiting factors”, first presented by Justus von Liebig in 1840 to explain the influence of chemical nutrients on plant growth. This approach can also be used to explain the existence of different ecosystems in different regions due to different climates [2]. The concept that the distribution of a species in any given region depends on the ecological niches available was introduced by Ellenberg, 1986, who designed ecograms for the tree species of Europe according to environmental factors [5,6].

The distribution of a species can be defined by abiotic factors, such as climate and substrate [7–10], and these factors were also used in models for the potential distribution of species. These models can be designed by a mechanistic or empirical approach as envelope models, climate response surfaces, homoclines, ecological niche models or species distribution models. Mechanistic models use critical climatic values for the physiological tolerance of different species to explain the range of distribution of the species. This means that a model for the distribution of a species is based on the critical physiological values measured in laboratory tests. In empirical models, climatic conditions are measured in the field together with the probability of occurrence of different species. These models result in a distribution pattern of the species and make it possible to predict the distribution under diverse climate scenarios. In contrast to mechanistic models, empirical models also show the biotic influences on the species [11]. An example of a mechanistic model for Northern Europe was STASH, which simulated the biogeographical distribution of individual species on a regional scale for the specific case of lodgepole pine (*Pinus contorta* Dougl.) in Scandinavia using scenarios from the Hamburg ECHAM3 climate model [12]. This bioclimatic model for the potential distributions of tree species in northern Europe was first published in 1996 [13].

Bioclimatic models on a global scale were used for modeling the carbon cycle. Models estimating the global net primary production (NPP, the difference between the CO₂ fixed in plant biomass and the CO₂ lost to the atmosphere) in units of bound carbon per unit area per annum are, for example, the Miami model [14], the Schuur model [15] and the National Center for Ecological Analysis and Synthesis (NCEAS) model [16]. The difference between these models is that the Schuur model was based on an extended data-set available for tropical regions, which allowed a better estimation under

tropical climatic conditions than the Miami model [15]. For the NCEAS model, datasets were available from the Ecosystem Model Data Intercomparison (EMDI) project, for both grass-dominated and tree-dominated landscapes. Thus it was possible to calculate correlations for both grass- and tree-dominated sites. Furthermore, besides the NPP, the above-ground NPP (ANPP) was estimated [16,17]. By comparing the coefficients of determination for the modeled and observed NPPs of the three models with each other, it was shown that the coefficient of the Miami model lies between the coefficients of the two other models [16]. Nevertheless, these models estimate the NPP and not the ABM (above-ground biomass production of a single species in terms of dry matter).

The NPP, which reflects the gross primary production (GPP) minus the atmospheric loss or respiration (R), can be split into above-ground NPP (ANPP) and below-ground NPP (BNPP). The NPP/GPP ratio is between 0.44 and 0.47, and the ratio ANPP/GPP about 0.29 [18,19], so that a combination of the two results in an ANPP/NPP ratio of 0.62 to 0.66. Furthermore, since a BNPP/NPP ratio of about 0.35 is known [18], it can be suggested that the ANPP/NPP ratio is one minus the BNPP/NPP ratio, which leads to an ANPP/NPP ratio of 0.65. These relations show that it should be possible to use a model developed to estimate the NPP as a basis for estimating the ABM.

When one of the NPP models introduced above is used to estimate the ABM, the limiting factors are climatic factors, particularly the mean annual temperature and the mean annual precipitation. This implies that soil conditions can reduce the growth rate and that fertilization supports an increase towards the maximum growth. This was demonstrated in Norway spruce (*Picea abies* Karst). The production of Norway spruce in northern Sweden increased by 300% and in southern Sweden by 100% when nutrition and water supply were no longer limiting factors [20]. This result impressively showed the impact of the soil, as well as the strong influence that nutrients have on plant production. For that reason and because of the relatively small differences and exceptions regarding soil conditions in Sweden [21,22], the soil was not considered in this study, and the focus was set on climatic factors. With regard to the climate, it is important to know which factors are considered and how the borders are defined. Temperature, for example, is a limiting factor not only during the growth season but also during the period of dormancy. *Miscanthus* spp. is a perennial C₄ grass, which has a photosynthetic mechanism obviously better adapted to low temperatures than that of various other C₄ plants [23]. Tests in Northern Europe showed, however, that especially in the first two years of cultivation, high losses are possible because of too low winter temperatures [24]. In this context, it can be assumed that climate is the limiting factor for distribution, while soil conditions have a strong impact on the growth rate.

Compared to models estimating the potential distribution of a species, the NPP models mentioned above for the production of all species on a site are defined only by the climate variables. A greater specification of a model for a single species has the advantage of providing a more detailed estimate of the productivity or distribution; but the disadvantage is that the model is valid for only one species. As these models are, in general, complex, the adaptation required for further species is quite extensive. In contrast, no method yet exists to transform a NPP model estimating the productivity over all species into a species-specific model estimating the ABM.

2. Objectives

The purpose of this work was to estimate and compare the potential above-ground biomass production of some selected species in terms of dry matter (ABM) on two specific sites in Sweden and to assess whether it is possible to transform a global NPP model—the Miami model—into a species-specific ABM model.

3. Materials and Methods

For modeling the productivity of a species based on climate, a combination of species-specific parameters and climatic factors must be used. The basis of this concept is to find a correlation between productivity and climatic factors and to transform this correlation into a mathematical function. For this work, the mathematical formulation of the Miami model was chosen as a basis; but, since the Miami model estimates NPP, some modifications had to be made. In addition to the transformation of NPP to above-ground biomass production in terms of dry matter (ABM) of a single species, these modifications include the effects of evapotranspiration (E) and species-specific climatic parameters (MAT_{nat}).

3.1. The Modified Miami Model

In the Miami model, the global net primary production (NPP) is estimated using the limiting productivity level defined either by the mean annual temperature in degrees centigrade (MAT) or the mean annual precipitation in millimeters (MAP). This means that NPP is expressed as:

$$NPP_{MAT} = 30 / (1 + e^{1.315 - 0.119MAT}) \text{ (Mg ha}^{-1}\text{a}^{-1}\text{)} \quad (1)$$

where MAT is the mean annual temperature ($^{\circ}\text{C}$), or as:

$$NPP_{MAP} = 30 \times (1 - e^{-0.00066MAP}) \text{ (Mg ha}^{-1}\text{a}^{-1}\text{)} \quad (2)$$

where MAP is the mean annual precipitation (mm). Each lower value of NPP is estimated from the two functions:

$$NPP = \text{Min}(NPP_{MAT}; NPP_{MAP}) \text{ (Mg ha}^{-1}\text{a}^{-1}\text{)} \quad (3)$$

Equations 1 and 2 are non-linear functions with a maximum productivity of 30 tons per hectare per year [14]. It is suggested that the NPP model can be transformed into a species-specific model by including two species-specific variables, the “maximum above-ground biomass production in the dry matter of a species” (BM_{max}) and the “MAT of the native distribution area of a species” (MAT_{nat}). Figure 1 gives an overview of the transformation of the Miami model. Moreover, it is supposed that evapotranspiration is linearly related to MAT [25].

The formulations of the two functions of the Miami model (Equations 1 and 2) support this type of modification. The exposed maxima of the two functions can be replaced by the BM_{max} found in the literature. As the Miami model was not designed for a single species, but for the collectivity of species on a single site, the BM_{max} has to be multiplied by a constant to avoid systematic underestimation of the biomass production of a single species. This constant is found by approximation of the modeled data to the literature data. Approximation during the validation process showed that for

the above-ground biomass production of a species in the dry matter (ABM) modeled by the MAT, multiplication of the BM_{max} by 1.5 (Equation 4) and for the ABM modeled by the MAP a multiplication of the MB_{max} by 2 (Equation 5) offer the best estimations by the model. The different BM_{max} potentials of single species leads to a fan of graphs differing only by a constant expressed by the differences in BM_{max} . This formulation is problematic for species with a high BM_{max} and adapted to warmer climates, such as *Paulownia* spp. [26,27]. Because of their weak frost resistance, such species can hardly survive under cooler conditions, and this leads to very low or no ABM at low MATs. This requirement can be fulfilled without changing the maximal values by extending the exponential MAT function (Equation 1) by MAT_{nat} within the exponent (Equation 4). The influence of MAT_{nat} expresses itself in an increasing gradient with an increase in MAT_{nat} , resulting in a shift of the zero-point in the direction of the inflexion point of the graph (Figure 2). This means that a higher MAT_{nat} leads to the start of biomass production at a higher temperature. The exact formulation of the term for MAT_{nat} was found by approximation during the validation process.

Figure 1. Illustration of the modification of the Miami model in order to transform the net primary production (NPP) model into a species-specific ABM model. Modification of the MAT function is shown on the left and of the MAP function on the right. MAT, mean annual temperature (°C); MAP, mean annual precipitation (mm); NPP, global net primary production ($Mg\ ha^{-1}a^{-1}$); MAT_{nat} , mean annual temperature of the native distribution area of a species (°C); ABM, modeled above-ground biomass production of a species in dry matter ($Mg\ ha^{-1}a^{-1}$); BM_{max} , maximal above-ground biomass production in dry matter of a species found in the literature ($Mg\ ha^{-1}a^{-1}$); E, mean annual evapotranspiration (mm).

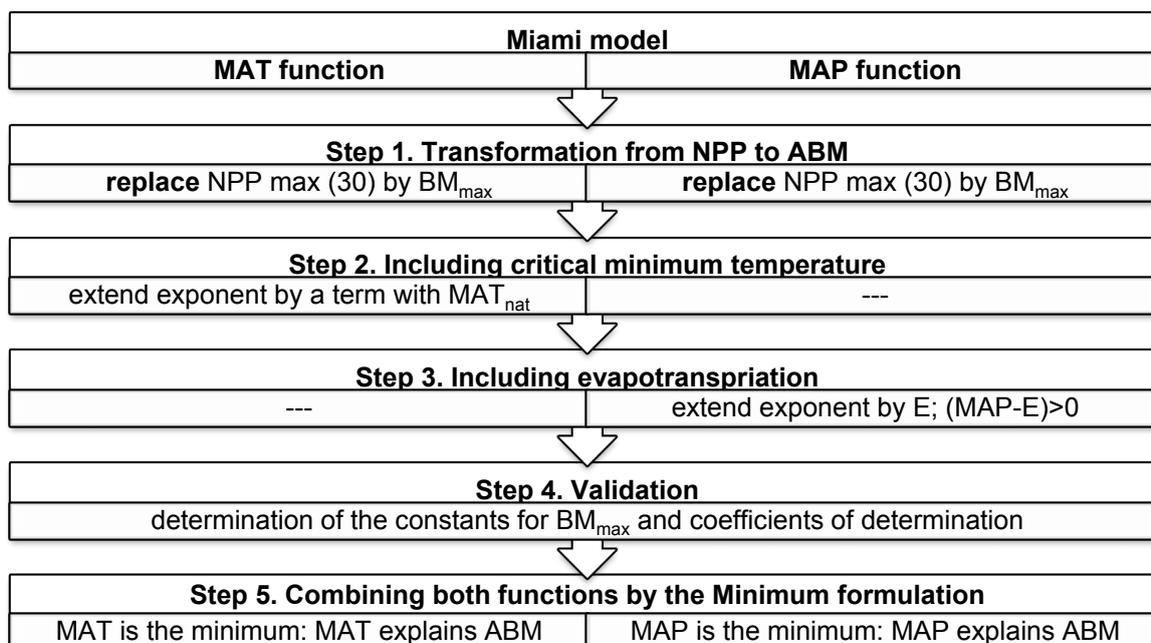
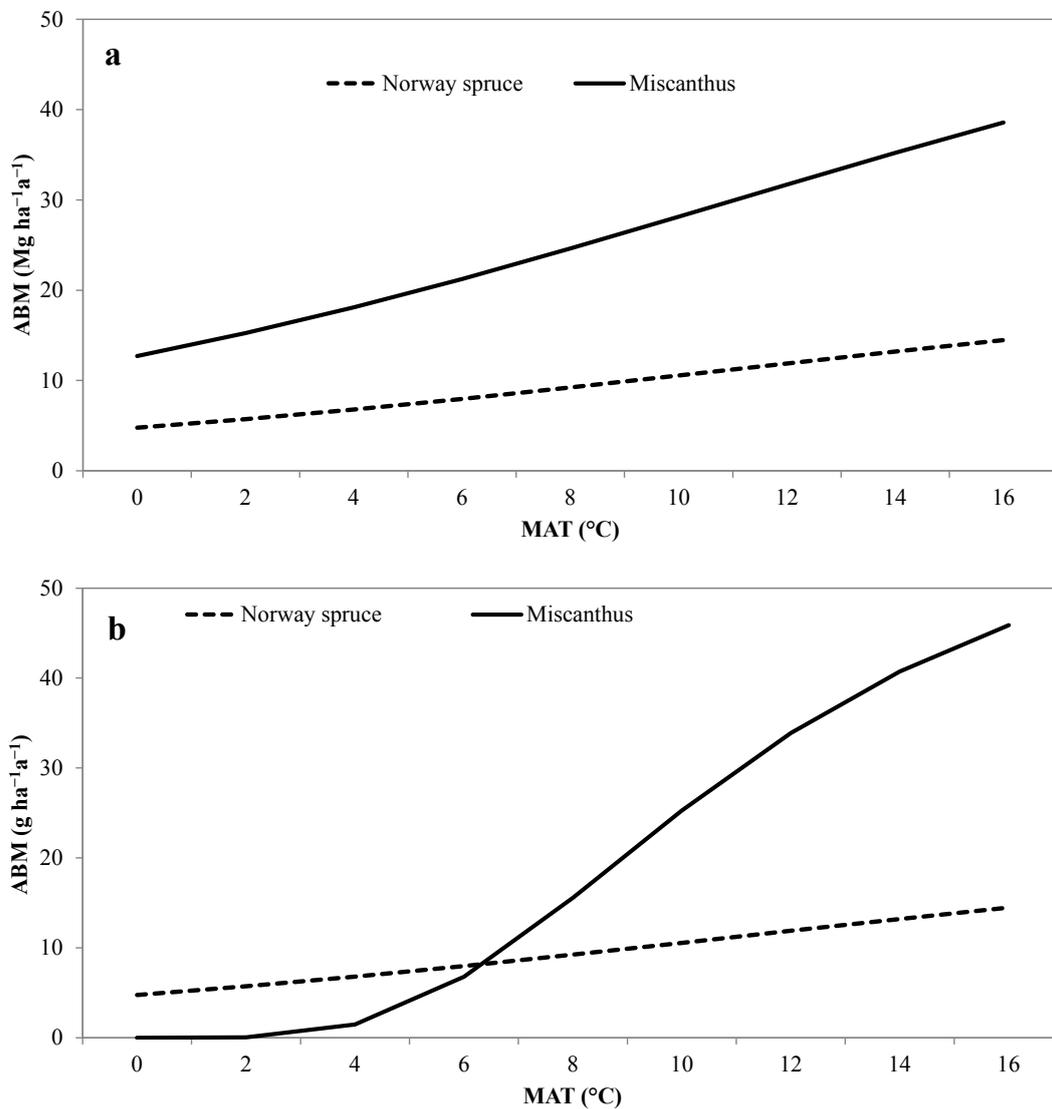
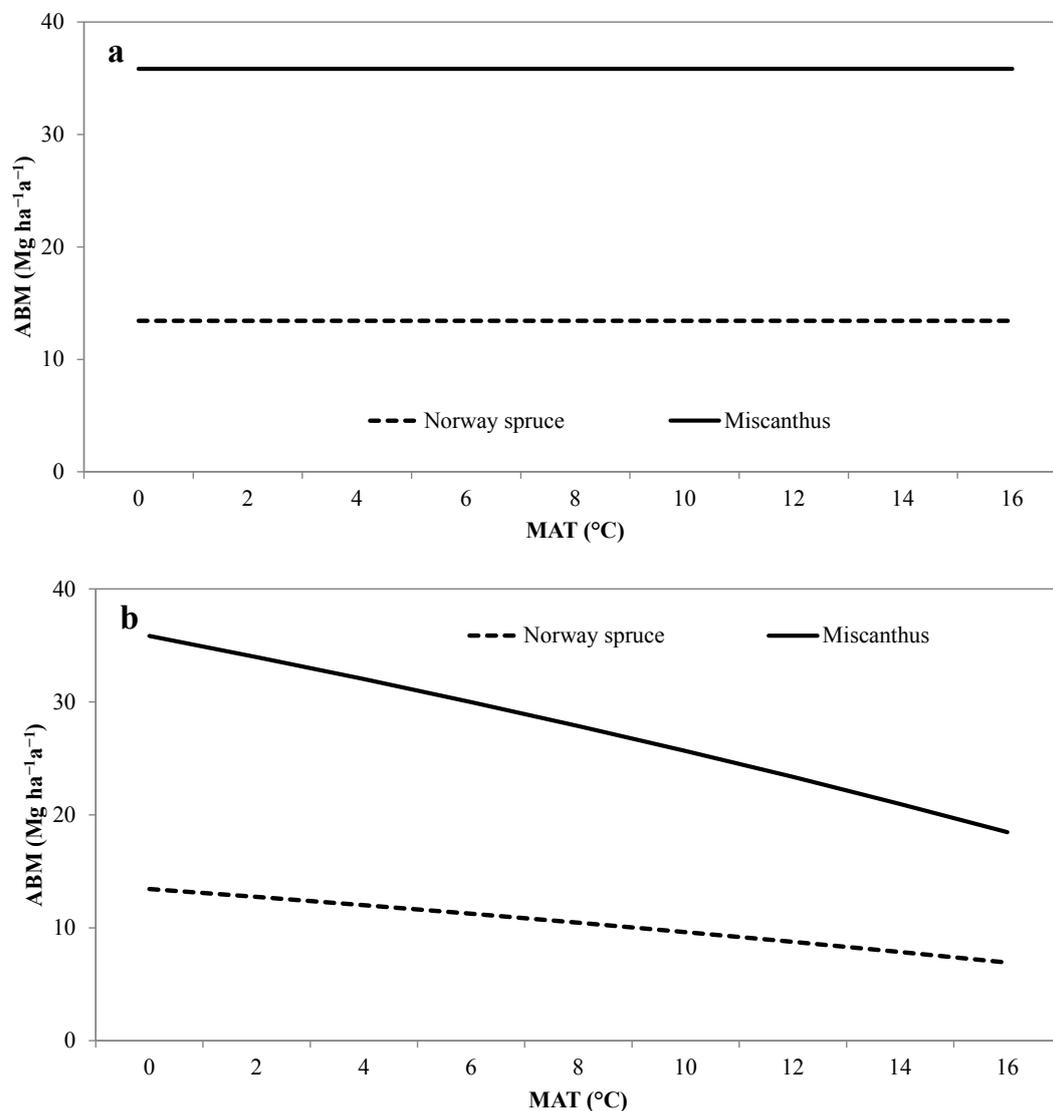


Figure 2. ABM modeled as a function of MAT using the MAT function and showing the influence of MAT_{nat} : **(a)** ABM of Norway spruce and *Miscanthus* without the influence of MAT_{nat} ; **(b)** Norway spruce with a MAT_{nat} of 1°C and *Miscanthus* with a MAT_{nat} of 18 °C; ABM, modeled above-ground biomass production of a species in dry matter ($Mg\ ha^{-1}a^{-1}$); MAT mean annual temperature (°C); MAT_{nat} , mean annual temperature of the native distribution area of a species (°C).



Higher temperatures may not always result in higher biomass production, as the evapotranspiration increases with increasing temperature [28]. This suggests that the ABM at a constant MAP estimated by the MAP function (Equation 2) is not constant over MAT, but rather decreases (Figure 3).

Figure 3. ABM modeled as a function of MAT using the MAP function with MAP = 900 mm and showing the influence of E: (a) ABM of Norway spruce and *Miscanthus* without E; and (b) after including E in the MAP function. ABM, modeled above-ground biomass production of a species in dry matter ($\text{Mg ha}^{-1}\text{a}^{-1}$); MAT, mean annual temperature ($^{\circ}\text{C}$); MAP, mean annual precipitation (mm); E, mean annual evapotranspiration (mm).



In addition, since the MAP equation is an exponential function, the mean annual evapotranspiration (E) has to be included in the exponent to achieve a change of gradient in the MAP graph (Equation 5). The linear function of E, as E represents the above-ground water loss compared to MAP as water input, suggests that the best formulation is to subtract E from MAP. A restriction in this formulation is that MAP minus E cannot be negative, as there is no negative biomass production. This restriction is solved by the logic formulation “if the term is negative, set it to zero”.

The final formulation of the modified Miami model is: the relationship between the modeled ABM and the MAT, including the MAT_{nat} for a single species on a specific site is given by:

$$\text{ABM}_{\text{MAT}} = 1.5 \times \text{BM}_{\text{max}} / (1 + e^{((\text{MAT} + \text{MAT}_{\text{nat}}) / (\text{MAT} + 1)) \times (1.315 - 0.119\text{MAT})}) \quad (\text{Mg ha}^{-1}\text{a}^{-1}) \quad (4)$$

In addition, the relationship between ABM and the MAP including the restriction that $E \leq \text{MAP}$ for a species on a specific site is given by:

$$\text{ABM}_{\text{MAP}} = 2 \times \text{BM}_{\text{max}} \times (1 - e^{-0.00066(\text{MAP} - E)}) \text{ (Mg ha}^{-1}\text{a}^{-1}\text{)} \quad (5)$$

The relationship between E and the MAT is given by:

$$E = \text{MAT} \times 31.4 \text{ (mm)} \quad (6)$$

The relationship between ABM, MAP and MAT is given by:

$$\text{ABM} = \text{Min} (\text{ABM}_{\text{MAT}}; \text{ABM}_{\text{MAP}}) \text{ (Mg ha}^{-1}\text{a}^{-1}\text{)} \quad (7)$$

Where BM_{max} is the maximum above-ground biomass production in the dry matter of a species ($\text{Mg ha}^{-1}\text{a}^{-1}$); MAT_{nat} is the mean annual temperature of the native distribution area of a species ($^{\circ}\text{C}$); E is the mean annual evapotranspiration (mm); MAT is the mean annual temperature ($^{\circ}\text{C}$); MAP is the mean annual precipitation (mm). Furthermore, the model has two restrictions regarding MAT that have to be considered:

- (1) $\text{MAT} \geq 0 \text{ }^{\circ}\text{C}$: this model gives an estimate of ABM, starting at a MAT of $0 \text{ }^{\circ}\text{C}$. It is not valid for values below $0 \text{ }^{\circ}\text{C}$.
- (2) $\text{MAT}_{\text{nat}} > 0 \text{ }^{\circ}\text{C}$: the MAT_{nat} must be positive because of the mathematical formulation.

3.2. Validation of the Modified Miami Model

For approximation and validation of the model, four different species were chosen, two of them adapted to cool climates, Norway spruce and reed canary grass (*Phalaris arundinacea* L.), and two highly productive species with a low frost resistance, *Paulownia* spp. and *Miscanthus* spp. Typical MAT_{nat} values and very high BM_{max} values for the four species were found in the literature as input data for the model. The modeled data (ABM) were compared (r^2) with the literature data (BM_{max}), and the result is presented in Table 1. The species-specific parameters were chosen as follows:

Norway spruce is a native species of taiga and montane forests, but cultivation has extended its area in Northern and Central Europe [29]. As the species is native to the boreal forest, a MAT_{nat} of $1 \text{ }^{\circ}\text{C}$ was suggested. The BM_{max} was chosen as $15 \text{ Mg ha}^{-1}\text{a}^{-1}$ in accordance with the findings of Bergh, Linderand Bergström [20].

Reed canary grass is native to nearly all of Europe, except above the Arctic Circle and some southern regions. The MAT_{nat} for reed canary grass was set at $2 \text{ }^{\circ}\text{C}$, as it grows best under cool and moist conditions [30], and the BM_{max} was chosen as $14 \text{ Mg ha}^{-1}\text{a}^{-1}$ [31].

Paulownia is a species originating from Asia with an average MAT_{nat} of about $17 \text{ }^{\circ}\text{C}$ [26]. While some species within this genus show frost resistance at about -20°C , others tolerate no frost [27]. Estimations based on growing tests gave values of 35 to 90 Mg ha^{-1} in two to three years [32]. For this study, $36 \text{ Mg ha}^{-1}\text{a}^{-1}$ (90 divided by 2.5) was chosen as BM_{max} .

Miscanthus is a highly productive species native to Southeast Asia with a MAT_{nat} of about $18 \text{ }^{\circ}\text{C}$ [23,33]. Growth tests on *Miscanthus* showed that yields of more than $40 \text{ Mg ha}^{-1}\text{a}^{-1}$ could be achieved [34].

Table 1. Literature data, parameters and output of the final modified Miami model. MAT, mean annual temperature; MAP, mean annual precipitation; BM_{max} , maximal above-ground biomass production of a species in dry matter; MAT_{nat} , mean annual temperature of the native distribution area of a species; ABM, modeled above-ground biomass production of a species in dry matter; r^2 , coefficient of determination for the final model.

Species	Literature data			Parameters		Output	r^2	References			
	MAT	MAP	BM_{max}	BM_{max}	MAT_{nat}	ABM					
	°C	mm	Mg ha ⁻¹ a ⁻¹	Mg ha ⁻¹ a ⁻¹	°C	Mg ha ⁻¹ a ⁻¹	-				
Norway spruce	0.0	718 *	4.5			4.8					
	2.0	718 *	6.2			5.7					
	4.0	718 *	7.9	15	1	6.8	0.90	[20,35–37]			
	6.0	718 *	9.6								
	8.0	718 *	11.3								
	10.0	718 *	13.0								
4.6	632	6.4							6.0		
5.9	750	6.2							7.0		
Reed canary grass	6.9	657	6.0	14	2	7.1	0.58	[31,38–42]			
	7.6	514	5.5								
	8.2	477	6.2								
	8.4	547	6.2								
	7.6	783	11.8							13.7	
Miscanthus	7.8	578	15.1			14.6					
	8.3	630	14.7			17.0					
	8.9	634	15.4			16.7					
	9.6	689	18.5			18.1					
	9.6	769	12.6	40	18	21.2	0.41	[23,24,34,43,44]			
	9.6	766	17.5								
	9.8	660	21.0								
	10.4	594	19.6								
	10.4	594	12.5								
	10.5	676	19.6								
	11.5	813	34.3								
									20.6		

* Estimated mean values.

Comparison of the BM_{max} values from the literature with the modeled BM_{max} shows an underestimation of BM_{max} , especially at higher MAT. The reason is that the BM_{max} values from the literature are extremely high values, partly achieved in growing tests. These extreme values can be much higher than average values, because of regional and experimental variation in nutrients and water supply. The difference increases with increasing MAT, because the importance of water as a limiting factor is intensified. The inclusion of MAP and E in the model therefore leads to an underestimation of BM_{max} by the model compared to the literature data. In the case of Norway spruce, for example, the high BM_{max} values were achieved by fertilization and irrigation. As modeling was based on a MAP of 718 mm, the impact of the water-limitation increased with increasing MAT.

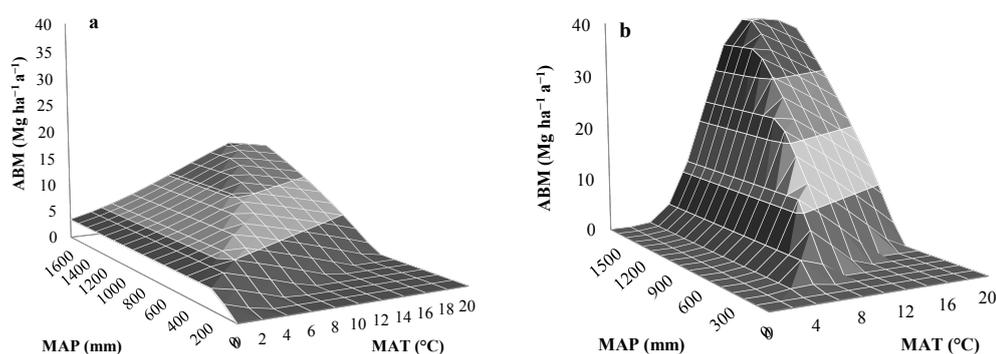
The values for the coefficient of determination (r^2) in Table 1 are between 0.41 and 0.90. Nevertheless, the small amount and small range of data may reduce the representative significance of the coefficient of determination. Therefore, the coefficient of determination for the different species, the paired samples correlation (t -test), was also used to increase the parameters by combining all the literature data presented in Table 1 under a single variable and comparing them to modeled data. For this purpose, the species-specific parameters (MAT_{nat} and BM_{max}) were exchanged either (1) by using the literature data directly as species-specific parameters or (2) by using average variables over all species for the parameters.

- (1) The literature data were compared to modeled data using each single MAT_{nat} and BM_{max} value in the literature directly as a species-specific parameter. The coefficient of determination of this modification was $r^2 = 0.903$ with $p < 0.0001$.
- (2) The literature data were compared to the modeled data using as parameters the average MAT_{nat} and BM_{max} values from the literature data. The coefficient of determination of this modification was $r^2 = 0.759$ with $p < 0.0001$.

For the paired sample comparisons (1) and (2), it could not be excluded that the two groups (the literature data and the modeled data) were equal, and this supports the hypothesis that the modified Miami model can be used to estimate the ABM. Thus, the results of the paired sample comparisons support the results of the validation based on the approximation process (Table 1), *i.e.*, that it is possible to estimate the ABM of a single species with the modified Miami model using species-specific variables.

To see the effect of the species-specific variables and the relation between the MAT (Equation 4) and MAP (Equation 5) functions (Equation 7), 3D graphs were drawn for Norway spruce and *Miscanthus* (Figure 4) using the species-specific parameters introduced above. The effect of the different BM_{max} and of the MAT_{nat} with a later start of biomass production in combination with a higher grade led to strong differences between the graphs for the two species.

Figure 4. ABM of (a) Norway spruce and (b) *Miscanthus* as functions of MAT and MAP; the species-specific parameters (BM_{max} , MAT_{nat}) result in different starts, different increases and, finally, in different amounts of ABM. ABM, modeled above-ground biomass production of a species in dry matter ($Mg\ ha^{-1}a^{-1}$); MAP, mean annual precipitation (mm); MAT, mean annual temperature ($^{\circ}C$); BM_{max} , maximal above-ground biomass production in dry matter of a species in the literature ($Mg\ ha^{-1}a^{-1}$); MAT_{nat} , mean annual temperature of the native distribution area of a species ($^{\circ}C$).



3.3. Choice of Species-Specific Parameters for the Modified MIAMI Model to Estimate the Above-Ground Biomass Production in Dry Matter for Different Selected Species

The modified Miami model was used to estimate the above-ground biomass production of a species in the dry matter (ABM) of different species on the two sites in Sweden. To compare the ABM of different species on different sites using the modified Miami model, the species-specific parameters for all the species and climate data for all the sites were needed. Table 2 shows the species-specific variables (BM_{max} and MAT_{nat}) in the literature for the different species. The climate data for the two sites are (1) for southern Sweden a MAT of 6 °C and MAP of 750 mm and (2) for northern Sweden a MAT of 3 °C and MAP of 601 mm. The site in northern Sweden lies in the fifth hardiness climate zone, and the site in southern Sweden in the seventh climate zone [45]. On both sites, three species, Norway spruce, Scots pine (*Pinus sylvestris* L.) and birch (*Betula* spp.), have together more than 90 percent of the standing volume of all land use classes [46]. In the region of the northern site, the total number of different tree species was 10. Of these, the most frequent was Scots pine with 43.1%vol. In the region of the southern site, sixteen species were found, the most frequent species being Norway spruce with 48.5%vol. The mean volume increment, including the growth of felled trees on forest land, was 3.0 and 5.7, respectively [46]. The soil on both sites is brown soil, and the dominant humus layers are duff and mull in the two regions. The regions are further characterized mainly by Podzols with high frequencies of clay [47]. For BM_{max} , the averages of the relatively higher biomass production values found in the literature were chosen. The MAT_{nat} values are either mean annual temperature values for the native region, when available, or averages of the mean annual temperature values from growing tests, plantations and forestry documented in the literature.

Table 2. Literature values of the maximum biomass production of a species in dry matter (BM_{max}) and the mean annual temperature of the native distribution area or the average mean annual temperature of the reported growing tests (MAT_{nat}) for different species.

Species	BM_{max}	MAT_{nat}	References
	Mg ha ⁻¹ a ⁻¹	°C	
<i>Picea abies</i> Karst	10.8	0.5	[48–53]
<i>Pinus sylvestris</i> L.	7.3	0.1	[50–55]
<i>Abies alba</i> Mill.	9.1	2.6	[50–57]
<i>Larix</i> spp.	9.1	3.9	[50–52,58–62]
<i>Pseudotsuga menziesii</i> Mirb. Franco	14.2	7.6	[52,63–66]
<i>Betula</i> spp.	6.7	0.3	[50,51,53,55,67]
<i>Alnus glutinosa</i> (L.) Gaertn.	7.7	2.7	[51,53,68]
<i>Populus nigra</i> L.	8.8	9.1	[69–75]
<i>Populus tremula</i> L.	8.8	1.9	[50,53,67,71,76–79]
<i>P. tremula</i> L. x <i>p. tremuloides</i> Michx.	11.4	1.0	[71,76,80,81]
<i>Salix</i> spp. (short rotation)	12.5	5.3	[50,71,74,75,82,83]
<i>Fraxinus excelsior</i> L.	8.7	9.6	[51,55,75,84–87]

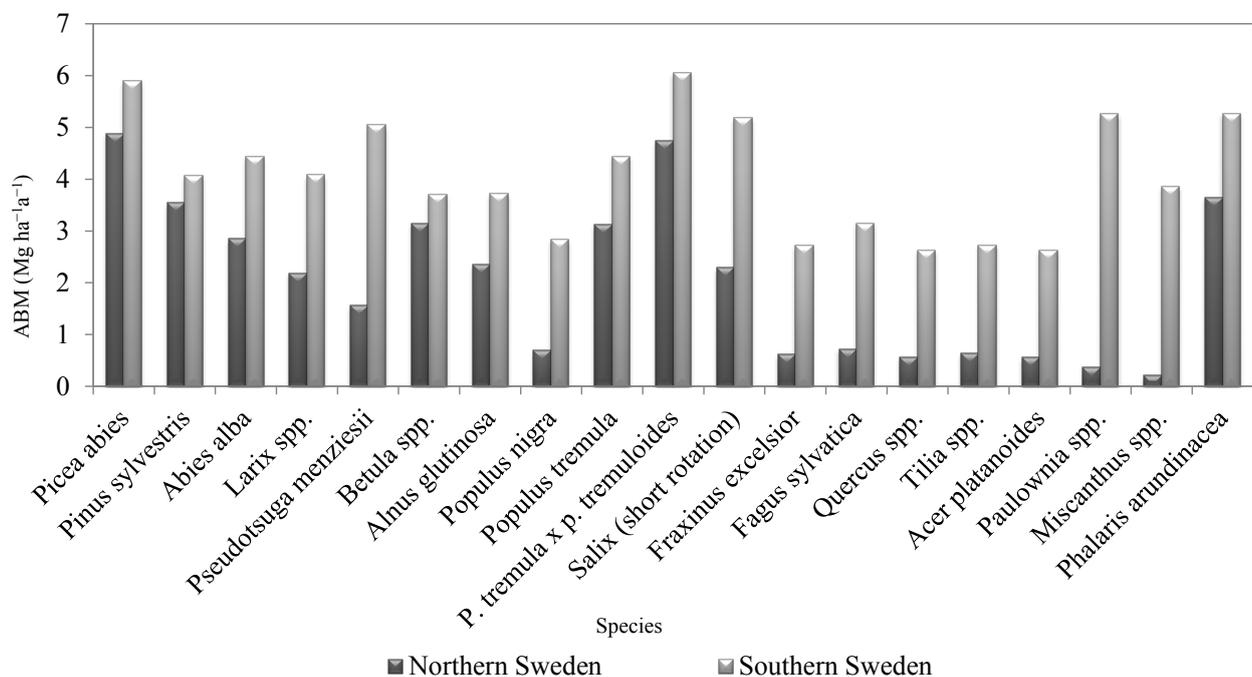
Table 2. Cont.

Species	BM _{max}	MAT _{nat}	References
	Mg ha ⁻¹ a ⁻¹	°C	
<i>Fagus sylvatica</i> L.	10.0	9.5	[51,52,55,73,84–87]
<i>Quercus</i> spp.	7.4	9.9	[51,52,73,75,84–86]
<i>Tilia</i> spp.	8.3	9.4	[84,88,89]
<i>Acer platanoides</i> L.	7.0	9.9	[51,75,84,85,87]
<i>Paulownia</i> spp.	28.3	16.8	[26,32,90,91]
<i>Miscanthus</i> spp.	23.5	18.4	[33,34,44,71,92,93]
<i>Phalaris arundinacea</i> L.	10.5	2.0	[31,41,42,94–96]

4. Results and Discussion

The species-specific variables presented in Table 2, together with the climate data for the two sites described above, were used as input parameters for the modified Miami model. Figure 5 shows the estimated above-ground biomass production of a species in dry matter (ABM) for the different species in both northern (black) and southern Sweden (grey). Although the absolute values for the different species may include a large variation in ABM for a single site, there is a clear trend in the differences between species at the two sites.

Figure 5. ABM of selected species for two sites in Sweden with a MAT of 3 °C and MAP of 601 mm (northern Sweden); and a MAT of 6 °C and MAP of 750 mm (southern Sweden); ABM, modeled above-ground biomass production of a species in dry matter (Mg ha⁻¹a⁻¹); MAT, mean annual temperature (°C); MAP, mean annual precipitation (mm).



Most of the species have ABM values in the range of 2 to 4 Mg ha⁻¹a⁻¹ for the site in southern Sweden (Figure 5 grey). Eight of the 19 species have an ABM between 4.5 and 6 Mg ha⁻¹a⁻¹. This is often less than half of the maximum biomass production presented in the literature (BM_{max}). The three species with the highest ABM are hybrid aspen (*P. tremula* L. × *P. tremuloides* Michx.), Norway spruce and reed canary grass.

In northern Sweden (Figure 5, black), the only species with a biomass production higher than 4 Mg ha⁻¹a⁻¹ are Norway spruce and hybrid aspen, while Scots pine, birch and aspen (*Populus tremula* L.) show a production of more than 3 Mg ha⁻¹a⁻¹. Reed canary grass also has a relatively high productivity, which explains why it is cultivated to such an extent in the boreal regions, a fact also supported by literature [38,40,95].

Of the broad-leaved trees, only the pioneer species, such as birch and aspen or alder (*Alnus glutinosa* (L.) Gaertn.), appear able to survive under boreal climatic conditions, which is also supported by forest inventory data [46]. The findings for other species are also supported by the literature: for *Paulownia* plantations, a MAT of at least 9 °C [97] or at least the sixth climate zone [98] are recommended, due to the low frost resistance of the fast growing variants of this species [26,27]; for *Miscanthus*, problems of cultivation in colder climate zones were reported [24]. Silver fir (*Abies alba* Mill.) and Douglas fir (*Pseudotsuga menziesii* (Mirb.) Franco) favour not overly cool climates with a relatively high precipitation [6,65,99,100]. For the species native to Sweden, the inventory data, such as the percentages of different species in the forest, support the results of the model [46].

The results, which include (1) a comparison of different species on one site and (2) a comparison of two different sites, show that it is possible to estimate the ABM for different species on different sites using the modified Miami model. This means that this model can be used as a decision support for the plantation of species under different climate conditions, because it shows that the extent of temperature or precipitation are limiting factors for biomass production. The results also demonstrate that both requirements, namely the variation in biomass production and the influence of the minimal temperature for the different species, can be fulfilled.

Nevertheless, it is only a basic formulation to be optimized and validated with a larger quantity of reference data. The most challenging task during the modeling and validation of the model was to find verified data. When data are found, they are generally presented in a more or less wide range, which gives a subjective perspective to the dataset used within the model, even if it was intended to be objective. Most critical for this concept is the use of MAT values from the native distribution area (MAT_{nat}) of a species. For some species, it is relatively easy to find plausible values, since they are well documented or native in a relatively limited area, but for others it is quite difficult and this means that mean values estimated over an extended area have to be used. Besides the inclusion of MAT_{nat}, the inclusion of evapotranspiration (E) in the MAP function is considered as critical. For this study, the linear correlation between MAT and E according to the findings of Komatsu, Maita and Otsuki [25] was used. However, since E was not allowed to exceed MAP, because this would result in negative biomass production, a logical function had to be included, setting any negative value to zero. Optimization might eliminate this problem. Furthermore, the model has not been verified for higher temperatures, where the MAP function becomes more important. According to the findings of Komatsu, Maita and Otsuki, verification in the range of higher temperatures will show whether it is

possible to use the function of E originally developed for mountain forest areas in Japan or whether further adaptation is necessary. However, the best option would be a global formulation for the correlation between MAT and E.

If optimization is carried out for water availability, attempts should be made to manage this via the distribution of precipitation over the year rather than soil water capacity; this approach would further exclude soil data. In general, soil texture and soil structure are critical impact factors, for instance because of water storage capacity and nutrient supply. In Sweden, soil data are quite similar over the entire country, and the factor soil could therefore be treated as a constant in the present study, but this would not be valid for other regions in Europe. On the other hand, topography, even on a small scale, also has a very strong influence on water and nutrient supply, and this would make more general modeling on a larger scale impossible. Furthermore, the exclusion of soil as a separate parameter increases the importance of the choice of the BM_{max} parameter. Even if soil with poorer conditions would be a reduction factor for the biomass production of the species, this fact should not influence their ranking. An overestimation of above-ground biomass production can be avoided by using average values for BM_{max} .

The advantage of this model is that the formulation is very easy. The two climate variables (MAT and MAP) necessary for estimating the biomass production for a site can be found quite easily in weather homepages. The challenge in this formulation is to find valid species-specific parameters (BM_{max} and MAT_{nat}). Apart from their effect on ABM, it is simple to demonstrate the influence of climate by changing the values for MAT and MAP.

The model provides some interesting results. Although some of the species listed in Table 2 have an enormous capacity for biomass production, it is not possible to use this potential in Sweden. *Paulownia* shows a relatively high ABM in southern Sweden, but this must be weighed against high risks, as this species grows on the border of its survival. The situation in northern Sweden seems to be quite different. If the foreign and hybrid species, in particular hybrid aspen, are excluded, only Norway spruce and reed canary grass have a biomass production higher than $4.5 \text{ Mg ha}^{-1}\text{a}^{-1}$, which explains and emphasizes the importance of Norway spruce for the forest industry in that region. Considering foreign species, Douglas fir, silver fir and larch (*Larix* spp.) might be alternatives in southern Sweden, but in northern Sweden hybrid aspen seems to be the only alternative tree species worth considering.

However, it must be noted that this simulation shows only average values for the above-ground biomass production of a species in dry matter (ABM) in the species-specific rotation period. Estimates of productivity in the developing phases of the plantation are not included, and that may result in a very low biomass production. Moreover, it must be considered that the fast-growing species differ from long-living species at a juvenile age in their periodical growth. Management aspects and the loss of biomass due to death have not been included.

5. Conclusions

A modified Miami model was used to estimate the above-ground biomass production of a species in dry matter (ABM) in tons per hectare per annum for different species. Modeling was done for two sites, one in northern and the other in southern Sweden. Results demonstrate that it is possible to compare

the biomass production of different species at one site, as well as the biomass production of different sites by modeling the ABM.

Results further illustrate the difficulty of introducing new, highly productive species in Sweden, especially in northern regions. In this study, the only tree species competitive with Norway spruce in northern Sweden in terms of the modeled ABM was hybrid aspen. The warmer climate in southern Sweden provides more opportunities for introducing new species, but since these new species presented in this paper would grow on the border of their distribution and survival area, the risks associated with cultivation are quite high. Other foreign species not included in this study might be more suitable for cultivation.

Conflicts of Interest

The authors declare no conflict of interest.

References

1. Lévêque, C. *Ecology: From Ecosystem to Biosphere*; BIOS Scientific Publishers Limited: Milton Park, UK, 2003; p. 472.
2. Nebel, B.J.; Wright, R.T. *Environmental Science: The Way the World Works*; Prentice Hall: Upper Saddle River, NJ, USA, 1993; p. 630.
3. Hoy, M.A. Classical biological control. In *Encyclopedia of Entomology*; Capinera, J.L., Ed.; Springer: Gainesville, FL, USA, 2008; Volume 2, pp. 906–923.
4. Orcutt, D.M.; Nilsen, E.T. *The Physiology of Plants under Stress: Soil and Biotic Factors*; Wiley: Hoboken, NJ, USA, 2000; Volume 2, p. 696.
5. Ellenberg, H. *Vegetation Mitteleuropas mit den Alpen: In ökologischer Sicht*; Ulmer: Stuttgart, Germany, 1986; p. 989.
6. Burschel, P.; Huss, J. *Grundriss des Waldbaus: Ein Leitfaden für Studium und Praxis*; Eugen Ulmer Verlag: Stuttgart, Germany, 1987; p. 487.
7. Akhtar, M.; Jaiswal, A.; Taj, G.; Jaiswal, J.P.; Qureshi, M.I.; Singh, N.K. DREB1/CBF transcription factors: Their structure, function and role in abiotic stress tolerance in plants. *J. Genet.* **2012**, *91*, 385–395.
8. Benito-Garzón, M.; Ruiz-Benito, P.; Zavala, M.A. Interspecific differences in tree growth and mortality responses to environmental drivers determine potential species distributional limits in Iberian forests. *Glob. Ecol. Biogeogr.* **2013**, *22*, 1141–1151.
9. Boucher-Lalonde, V.; Morin, A.; Currie, D.J. How are tree species distributed in climatic space? A simple and general pattern. *Glob. Ecol. Biogeogr.* **2012**, *21*, 1157–1166.
10. Pellissier, L.; Anne Bråthen, K.; Pottier, J.; Randin, C.F.; Vittoz, P.; Dubuis, A.; Yoccoz, N.G.; Alm, T.; Zimmermann, N.E.; Guisan, A.; *et al.* Species distribution models reveal apparent competitive and facilitative effects of a dominant species on the distribution of tundra plants. *Ecography* **2010**, *33*, 1004–1014.
11. Jeschke, J.M.; Strayer, D.L. Usefulness of bioclimatic models for studying climate change and invasive species. In *The Year in Ecology and Conservation Biology*; Ostfeld, R.S., Schlesinger, W.H., Eds.; Blackwell Scientific Publishing: Boston, MA, USA, 2008; p. 1134.

12. Sykes, M.T. Modelling the potential distribution and community dynamics of lodgepole pine (*Pinus contorta* Dougl. ex. Loud.) in Scandinavia. *For. Ecol. Manag.* **2001**, *141*, 69–84.
13. Sykes, M.T.; Prentice, I.C.; Cramer, W. A bioclimatic model for the potential distributions of north European tree species under present and future climates. *J. Biogeogr.* **1996**, *23*, 203–233.
14. Lieth, H. Modeling the primary productivity of the world. In *Primary Productivity of the Biosphere*; Lieth, H., Whittaker, R.H., Eds.; Springer-Verlag: New York, NY, USA, 1975; pp. 237–263.
15. Schuur, E.A.G. Productivity and global climate revisited: The sensitivity of tropical forest growth to precipitation. *Ecology* **2003**, *84*, 1165–1170.
16. Grosso, S.D.; Parton, W.; Stohlgren, T.; Zheng, D.; Bachelet, D.; Prince, S.; Hibbard, K.; Olson, R. Global potential net primary production predicted from vegetation class, precipitation, and temperature. *Ecology* **2008**, *89*, 2117–2126.
17. Shoo, L.P.; Ramirez, V.V. Global potential net primary production predicted from vegetation class, precipitation, and temperature: Comment. *Ecology* **2010**, *91*, 921–923.
18. Zha, T.S.; Barr, A.G.; Bernier, P.Y.; Lavigne, M.B.; Trofymow, J.A.; Amiro, B.D.; Arain, M.A.; Bhatti, J.S.; Black, T.A.; Margolis, H.A.; *et al.* Gross and aboveground net primary production at Canadian forest carbon flux sites. *Agric. For. Meteorol.* **2013**, *174–175*, 54–64.
19. Waring, R.H.; Landsberg, J.J.; Williams, M. Net primary production of forests: A constant fraction of gross primary production? *Tree Physiol.* **1998**, *18*, 129–134.
20. Bergh, J.; Linder, S.; Bergström, J. Potential production of Norway spruce in Sweden. *For. Ecol. Manag.* **2005**, *204*, 1–10.
21. Troedsson, T.; Nykvist, N. *Marklära och Markvård*; Almqvist & Wiksell: Stockholm, Sweden, 1973; p. 403.
22. Troedsson, T. *Mark, Marksligtage och Markhushållning*; LTs förlag: Helsingborg, Sweden, 1975; p. 167.
23. McKervey, Z.; Woods, V.B.; Easson, D.L. *Miscanthus as an Energy Crop and Its Potential for Northern Ireland: A Review of Current Knowledge*; Agri-Food and Biosciences Institute, Global Research Unit: Hillsborough, UK, 2008; p. 72.
24. Greef, J.M.; Deuter, M.; Jung, C.; Schondelmaier, J. Genetic diversity of European *Miscanthus* species revealed by AFLP fingerprinting. *Genet. Resour. Crop Evol.* **1997**, *44*, 185–195.
25. Komatsu, H.; Maita, E.; Otsuki, K. A model to estimate annual forest evapotranspiration in Japan from mean annual temperature. *J. Hydrol.* **2008**, *348*, 330–340.
26. Zhu, Z.-H.; Chao, C.-J.; Lu, X.-Y.; Xiong, Y.G. *Paulownia in China: Cultivation and Utilization*; Chinese Academy of Forestry/Asian Network for Biological Sciences/International Development Research Center: Beijing, China, 1986; p. 65.
27. Schröder, B. Paulownien für die Plantagenwirtschaft—Charakteristikum der Paulownia Sorten. Available online: <http://www.paulownia-baumschule.de/forstpflanzen/> (accessed on 12 April 2013).
28. Helfer, F.; Lemckert, C.; Zhang, H. Impacts of climate change on temperature and evaporation from a large reservoir in Australia. *J. Hydrol.* **2012**, *475*, 365–378.

29. Spiecker, H. The growth of Norway spruce (*Picea abies* [L.] Karst.) in Europe within and beyond its natural range. In *Forest Ecosystem Restoration*, Proceedings of the International Conference, Vienna, Austria, 10–12 April 2000; Hasenauer, H., Ed.; Institute of Forest Growth Research: Wien, Austria, 2000; pp. 247–256.
30. Lavergne, S.; Molofsky, J. Reed canary grass (*Phalaris arundinacea*) as a biological model in the study of plant invasions. *Crit. Rev. Plant Sci.* **2004**, *23*, 415–429.
31. Stražil, Z. Evaluation of reed canary grass (*Phalaris arundinacea* L.) grown for energy use. *Res. Agric. Eng.* **2012**, *58*, 119–130.
32. Lawrence, J.S. *TGG Paulownia Biomass Production*; Toad Gully Growers: Melbourne, Australia, 2011; p. 2.
33. Nishiwaki, A.; Mizuguti, A.; Kuwabara, S.; Toma, Y.; Ishigaki, G.; Miyashita, T.; Yamada, T.; Matuura, H.; Yamaguchi, S.; Rayburn, A.L.; *et al.* Discovery of natural miscanthus (*Poaceae*) triploid plants in sympatric populations of *Miscanthus sacchariflorus* and *Miscanthus sinensis* in southern Japan. *Am. J. Bot.* **2011**, *98*, 154–159.
34. Clifton-Brown, J.C.; Lewandowski, I.; Andersson, B.; Basch, G.; Christian, D.G.; Kjeldsen, J.B.; Jørgensen, U.; Mortensen, J.V.; Riche, A.B.; Schwarz, K.-U.; *et al.* Performance of 15 miscanthus genotypes at five sites in Europe. *Agron. J.* **2001**, *93*, 1013–1019.
35. Schulze, E.D. Carbon and nitrogen cycling in European forest ecosystems. In *Ecological Studies*; Schulze, E.D., Ed.; Springer-Verlag: Heidelberg, Germany, 2000; Volume 142, p. 3–14.
36. *Virkesförrådet Fördelat på Trädslag Inom Diameterklasser. Exkl Torra och Vindfällda Träd. Alla Ågoslag. 2006–2010*; SLU (Institutionen för skoglig resurshushållning), Sveriges officiella statistik: Umeå, Sweden, 2011; p 128.
37. SMHI Årsmedelnederbörden sedan 1860. Available online: <http://www.smhi.se/klimatdata/meteorologi/nederbord/1.2887> (accessed on 22 March 2013).
38. Xiong, S.; Lötjönen, T.; Knuutila, K. *Energiproduktion. från Rörflen: Handbok för el och Värmeproduktion (Energy Conversion with Reed Canary Grass: A Guide for Electricity and Heat Production)*; Swedish University of Agriculture, Unit of Biomass Technology and Chemistry: Umeå, Sweden, 2008; p. 36.
39. Carlsson, E. *Rörflen som Alternativt Strukturfoder och Strömedel—Är Rörflen ett Ekonomiskt Alternativ som Strukturfoder och Strömedel?* Sveriges Lantbruksuniversitet, Fakulteten för Lantskapsplanering, Trädgårds-och Jordbruksvetenskap: Alnarp, Sweden, 2012; p. 25.
40. Pahkala, K.; Aalto, M.; Isolahti, M.; Poikola, J.; Jauhiainen, L. Large-scale energy grass farming for power plants—A case study from Ostrobothnia, Finland. *Biomass Bioenergy* **2008**, *32*, 1009–1015.
41. Pahkala, K.; Pihala, M. Different plant parts as raw material for fuel and pulp production. *Ind. Crops Prod.* **2000**, *11*, 119–128.
42. Jansone, B.; Rancane, S.; Berzins, P.; Stesele, V. Reed canary grass (*Phalaris Arundinacea* L.) in natural biocenosis of Latvia, research experiments and production fields. In Proceedings of the International Scientific Conference: Renewable Energy and Energy Efficiency, Jelgava, Latvia, 28–30 May 2012; Latvia University of Agriculture: Jelgava, Latvia, 2012; pp. 61–65.

43. Lewandowski, I.; Clifton-Brown, J.C.; Andersson, B.; Basch, G.; Christian, D.G.; Jørgensen, U.; Jones, M.B.; Riche, A.B.; Schwarz, K.U.; Tayebi, K.; *et al.* Environment and harvest time affects the combustion qualities of miscanthus genotypes. *Agron. J.* **2003**, *95*, 1274–1280.
44. Clifton-Brown, J.C.; Stampfl, P.F.; Jones, M.B. Miscanthus biomass production for energy in Europe and its potential contribution to decreasing fossil fuel carbon emissions. *Glob. Chang. Biol.* **2004**, *10*, 509–518.
45. Schulz, B. Average annual minimum temperatures for Europe. In *Flora der Gehölze: Bestimmung, Eigenschaften, Verwendung*; Roloff, A., Bärtels, A., Eds.; Ulmer Eugen Verlag: Stuttgart, Germany, 2008; p. 1.
46. *Skogsdata 2013*; SLU: Umeå, Sweden, 2013; p. 157.
47. Troedsson, T.; Wiberg, M. *Sveriges Jordmåner (Soil Map of Sweden)*; Sveriges Lantbruksuniversitetet, Institutionen för Skoglig Marklära: Uppsala, Sweden, 1986; p. 2.
48. Bergh, J.; Linder, S.; Lundmark, T.; Elfving, B. The effect of water and nutrient availability on the productivity of Norway spruce in northern and southern Sweden. *For. Ecol. Manag.* **1999**, *119*, 51–62.
49. Johansson, T. Biomass production of Norway spruce (*Picea abies* (L.) Karst.) growing on abandoned farmland. *Silva Fenn.* **1999**, *33*, 261–280.
50. Lässig, R.; Močalov, S.A. Frequency and characteristics of severe storms in the Urals and their influence on the development, structure and management of the boreal forests. *For. Ecol. Manag.* **2000**, *135*, 179–194.
51. Kramer, H.; Gussone, H.A.; Schober, R. *Waldwachstumslehre: Ökologische und Anthropogene Einflüsse auf das Wachstum des Waldes, Seine Massen—Und Wertleistung und die Bestandessicherheit*; mit 165 Tabellen; Paul Parey: Hamburg, Germany, 1988; p. 374.
52. Kaltschmitt, M.; Hartmann, H.; Hofbauer, H. *Energie aus Biomasse: Grundlagen, Techniken und Verfahren*; Springer: Heidelberg, Germany, 2009; p. 1032.
53. Schulze, E.D.; Vygodskaya, N.N.; Tchepakova, N.M.; Czimeczik, C.I.; Kozlov, D.N.; Lloyd, J.; Mollicone, D.; Parfenova, E.; Sidorov, K.N.; Varlagin, A.V.; *et al.* The Eurosiberian Transect: An introduction to the experimental region. *Tellus B* **2002**, *54*, 421–428.
54. Xiao, C.-W.; Yuste, J.C.; Janssens, I.A.; Roskams, P.; Nachtergale, L.; Carrara, A.; Sanchez, B.Y.; Ceulemans, R. Above- and belowground biomass and net primary production in a 73-year-old Scots pine forest. *Tree Physiol.* **2003**, *23*, 505–516.
55. Bray, J.R. Root production and the estimation of net productivity. *Can. J. Bot.* **1963**, *41*, 65–72.
56. Becker, M. The role of climate on present and past vitality of silver fir forests in the Vosges mountains of northeastern France. *Can. J. For. Res.* **1989**, *19*, 1110–1117.
57. Macias, M.; Andreu, L.; Bosch, O.; Camarero, J.J.; Gutiérrez, E. Increasing aridity is enhancing Silver fir (*Abies Alba* Mill.) water stress in its south-western distribution limit. *Clim. Chang.* **2006**, *79*, 289–313.
58. Zavitkovski, J.; Strong, T.F. *Biomass Production of 12-Year-Old Intensively Cultured Larix Eurolepis*; North Central Forest Experiment Station, Forest Service, USDA: St. Paul, MN, USA, 1984; p. 3.

59. Peeters, A.; Vanbellinghen, C.; Frame, J. *Wild and Sown Grasses: Profiles of a Temperate Species Selection: Ecology, Biodiversity and Use*; Food and Agriculture Organization of the United Nations/Wiley-Blackwell: Rome, Italy, 2004; p. 328.
60. Fior, C.; Anfodillo, T.; Masotto, G.P.e.A. *Biomass Partitioning and Structural Traits in Larix decidua Mill. and Pinus Cembra L. at Different Elevation*; Treeline Ecology Research Unit, Dipartimento Territorio e Sistemi Agro Forestali, Università di Padova: Legnaro, Italy, 2005; p. 11.
61. Bollschweiler, M.; Stoffel, M.; Schneuwly, D.M.; Bourqui, K. Traumatic resin ducts in *Larix decidua* stems impacted by debris flows. *Tree Physiol.* **2008**, *28*, 255–263.
62. Lingua, E.; Cherubini, P.; Motta, R.; Nola, P. Spatial structure along an altitudinal gradient in the Italian central Alps suggests competition and facilitation among coniferous species. *J. Veg. Sci.* **2008**, *19*, 425–436.
63. Grier, C.C.; Logan, R.S. Old-growth *Pseudotsuga menziesii* communities of a Western Oregon watershed: Biomass distribution and production budgets. *Ecol. Monogr.* **1977**, *47*, 373–400.
64. Kohnle, U. Douglasienanbau in Südwest-Deutschland: Waldbauliche Erfolgsfaktoren. In *BFW-Praxisinformationen*; Mauser, H., Ed.; Bundesforschungs- und Ausbildungszentrum für Wald, Naturgefahren und Landschaft (BFW): Wien, Austria, 2007; Volume 16, pp. 12–13.
65. Hermann, R.K.; Lavender, D.P. *Pseudotsuga menziesii* (Mirb.) Franco. Douglas-Fir. Available online: http://www.na.fs.fed.us/pubs/silvics_manual/Volume_1/pseudotsuga/menziesii.htm (accessed on 17 May 2013).
66. Zhang, Q.-B.; Hebda, R.J. Variation in radial growth patterns of *Pseudotsuga menziesii* on the central coast of British Columbia, Canada. *Can. J. For. Res.* **2004**, *34*, 1946–1954.
67. Syrjanen, K.; Kalliola, R.; Puolasmaa, A.; Mattsson, J. Landscape structure and forest dynamics in subcontinental russian european taiga. *Ann. Zool. Fenn.* **1994**, *31*, 19–34.
68. Johansson, T. Biomass equations for determining fractions of common and grey alders growing on abandoned farmland and some practical implications. *Biomass Bioenergy* **2000**, *18*, 147–159.
69. Laureysens, I.; Deraedt, W.; Indehherberge, T.; Ceulemans, R. Population dynamics in a 6-year old coppice culture of poplar. I. Clonal differences in stool mortality, shoot dynamics and shoot diameter distribution in relation to biomass production. *Biomass Bioenergy* **2003**, *24*, 81–95.
70. Trnka, M.; Fialová, J.; Koutecký, V.; Fajman, M.; Žalud, Z.; Hejduk, S. Biomass production and survival rates of selected poplar clones grown under a short-rotation system on arable land. *Plant Soil Environ.* **2008**, *54*, 78–88.
71. Diamantidis, N.D.; Koukios, E.G. Agricultural crops and residues as feedstocks for non-food products in Western Europe. *Ind. Crops Prod.* **2000**, *11*, 97–106.
72. Benetka, V.; Bartáková, I.; Mottl, J. Productivity of *Populus nigra* L. ssp. *nigra* under short-rotation culture in marginal areas. *Biomass Bioenergy* **2002**, *23*, 327–336.
73. Helle, G.; Schleser, G.H. Beyond CO₂-fixation by Rubisco—An interpretation of 13C/12C variations in tree rings from novel intra-seasonal studies on broad-leaf trees. *Plant Cell Environ.* **2004**, *27*, 367–380.
74. Barsoum, N. Relative contributions of sexual and asexual regeneration strategies in *Populus nigra* and *Salix alba* during the first years of establishment on a braided gravel bed river. *Evol. Ecol.* **2001**, *15*, 255–279.

75. Deiller, A.-F.; Walter, J.-M.N.; Trémolières, M. Effects of flood interruption on species richness, diversity and floristic composition of woody regeneration in the Upper Rhine alluvial hardwood forest. *Regul. Rivers: Res. Manag.* **2001**, *17*, 393–405.
76. Liesebach, M.; von Wuehlisch, G.; Muhs, H.J. Aspen for short-rotation coppice plantations on agricultural sites in Germany: Effects of spacing and rotation time on growth and biomass production of aspen progenies. *For. Ecol. Manag.* **1999**, *121*, 25–39.
77. Johansson, T. Biomass equations for determining fractions of European aspen growing on abandoned farmland and some practical implications. *Biomass Bioenergy* **1999**, *17*, 471–480.
78. Latva-Karjanmaa, T.; Suvanto, L.; Leinonen, K.; Rita, H. Emergence and survival of *Populus tremula* seedlings under varying moisture conditions. *Can. J. For. Res.* **2003**, *33*, 2081–2088.
79. Ellis, C.J.; Coppins, B.J. Reproductive strategy and the compositional dynamics of crustose lichen communities on aspen (*Populus tremula* L.) in Scotland. *Lichenologist* **2007**, *39*, 377–391.
80. Ilstedt, B.; Gullberg, U. Genetic variation in a 26-year old hybrid aspen trial in southern Sweden. *Scand. J. For. Res.* **1993**, *8*, 185–192.
81. Christersson, L. Wood production potential in poplar plantations in Sweden. *Biomass Bioenergy* **2010**, *34*, 1289–1299.
82. Wilkinson, J.M.; Evans, E.J.; Bilsborrow, P.E.; Wright, C.; Hewison, W.O.; Pilbeam, D.J. Yield of willow cultivars at different planting densities in a commercial short rotation coppice in the north of England. *Biomass Bioenergy* **2007**, *31*, 469–474.
83. Buček, A.; Maděra, P.; Úradníček, L. Czech Approach to Implementation of Ecological Network. *J. Landsc. Ecol.* **2012**, *5*, 14–28.
84. Neiryneck, J.; Mirtcheva, S.; Sioen, G.; Lust, N. Impact of *Tilia platyphyllos* Scop., *Fraxinus excelsior* L., *Acer pseudoplatanus* L., *Quercus robur* L. and *Fagus sylvatica* L. on earthworm biomass and physico-chemical properties of a loamy topsoil. *For. Ecol. Manag.* **2000**, *133*, 275–286.
85. Vitasse, Y.; Porté, A.; Kremer, A.; Michalet, R.; Delzon, S. Responses of canopy duration to temperature changes in four temperate tree species: Relative contributions of spring and autumn leaf phenology. *Oecologia* **2009**, *161*, 187–198.
86. Ritter, E.; Dalsgaard, L.; Einhorn, K.S. Light, temperature and soil moisture regimes following gap formation in a semi-natural beech-dominated forest in Denmark. *For. Ecol. Manag.* **2005**, *206*, 15–33.
87. Petritan, A.M.; von Lüpke, B.; Petritan, I.C. Effects of shade on growth and mortality of maple (*Acer pseudoplatanus*), ash (*Fraxinus excelsior*) and beech (*Fagus sylvatica*) saplings. *Forestry* **2007**, *80*, 397–412.
88. Radoglou, K.; Dobrowolska, D.; Spyroglou, G.; Nicolescu, V.N. A review on the ecology and silviculture of limes (*Tilia cordata* Mill., *Tilia platyphyllos* Scop. and *Tilia tomentosa* Moench.) in Europe. *Bodenkultur-Wien and München* **2009**, *60*, 9–20.
89. Kirlum, F. *Vegetationskundliche Untersuchungen im “Colbitzer Lindenwald”*; FH-Eberswalde: Eberswalde, Germany, 1995; p. 35.
90. López, F.; Pérez, A.; Zamudio, M.A.M.; de Alva, H.E.; García, J.C. Paulownia as raw material for solid biofuel and cellulose pulp. *Biomass Bioenergy* **2012**, *45*, 77–86.

91. Cokesa, V.; Koprivica, M.; Markovic, N.; Stajic, S. Production experiments of Paulownia in Vojvodina. In Proceedings of an International Scientific Conference Marking 75 Years of the Forest Research Institute of the Bulgarian Academy of Sciences, Sofia, Bulgaria, 1–5 October 2003; Rossnev, B., Kitanova, S., Alexandrov, A., Raev, I., Tsakov, H., Dimitrov, V., Grozeva, M., Petrova, R., Popov, G., Grigorov, G., Eds.; Forest Research Institute: Sofia, Bulgaria, 2004; Volume 1, pp. 248–255.
92. Hodgson, E.M.; Nowakowski, D.J.; Shield, I.; Riche, A.; Bridgwater, A.V.; Clifton-Brown, J.C.; Donnison, I.S. Variation in miscanthus chemical composition and implications for conversion by pyrolysis and thermo-chemical bio-refining for fuels and chemicals. *Bioresour. Technol.* **2011**, *102*, 3411–3418.
93. Heaton, E.A.; Dohleman, F.G.; Long, S.P. Meeting US biofuel goals with less land: The potential of Miscanthus. *Glob. Chang. Biol.* **2008**, *14*, 2000–2014.
94. Greenhalf, C.E.; Nowakowski, D.J.; Bridgwater, A.V.; Titiloye, J.; Yates, N.; Riche, A.; Shield, I. Thermochemical characterisation of straws and high yielding perennial grasses. *Ind. Crops Prod.* **2012**, *36*, 449–459.
95. Finell, M. The Use of Reed Canary Grass (*Phalaris arundinacea*) as a Short Fibre Raw Material for the Pulp and Paper Industry. Ph.D. Thesis, Swedish University of Agricultural Sciences, Umeå, Sweden, 2003; Volume 424, p. 53.
96. Casler, M.; Cherney, J.; Brummer, E.C. Biomass yield of naturalized populations and cultivars of reed canary grass. *Bioenergy Res.* **2009**, *2*, 165–173.
97. Paulownia-welt.de Paulownia. Available online: <http://www.paulownia-welt.de/paulownia/index.html> (accessed on 12 April 2013).
98. Institute_LL_C Paulownia growing zones. Available online: <http://www.worldpaulownia.com/html/zones.html> (accessed on 12 April 2013).
99. Bijak, S. Tree-ring chronology of Silver fir and its dependence on climate of the Kaszubskie lakeland (northern Poland). *Geochromonetria* **2010**, *35*, 91–94.
100. Peguero-Pina, J.J.; Camarero, J.J.; Abadía, A.; Martín, E.; González-Cascón, R.; Morales, F.; Gil-Peigrín, E. Physiological performance of silver-fir (*Abies alba* Mill.) populations under contrasting climates near the south-western distribution limit of the species. *Flora Morphol. Distrib. Funct. Ecol. Plants* **2007**, *202*, 226–236.