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Assessing Soil Metal Levels in an Industrial Environment of Northwestern China and the Phytoremediation Potential of Its Native Plants

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Abstract: Various industrial activities contribute heavy metals to terrestrial ecosystems. In order to evaluate the soil quality of industrial areas and to identify the potential phytoremediator from the native plant species, we collected 45 surface soil samples and 21 plant species in a typical industrial area of northwestern China. The results showed that the average values of the Cd, Cr, As, Pb, Cu, and Zn in the soils were 36.91, 1.67, 7.20, 1.38, 1.27, and 6.66 times, respectively, compared with the corresponding background values. The average single factor pollution index for heavy metals decreased in the order of Cd > As > Zn > Cr > Cu > Pb. The study area was seriously polluted by Cd and As, slightly polluted by Zn, and had relatively little contamination by Cr, Pb, and Cu. In terms of the average Nemerow synthetic pollution index in every sampling site, 97.78% of the samples were seriously polluted and 2.22% of the samples were moderately polluted, which indicated that almost all of the samples in the industrial area were seriously polluted. The results of the biomass, heavy metal concentrations, bioconcentration factors (BCF), and translocation factors (TF) for the native plants showed that Achnatherum splendens for metal Cr presented a phytostabilization potential, Artemisia scoparia and Echinochloa crusgalli for metal Cu and Halogeton arachnoideus for metal Zn presented a phytoextraction potential, and all of the studied plants were limited as phytoremediators for Cd or Pb contaminated soil.

Keywords: industrial area; heavy metal pollution; pollution assessment; dominant plant; potential phytoremediator

1. Introduction

Heavy metal pollution has become a worldwide environmental concern because of its latency, toxicity, and contamination within soils over time [1]. Heavy metal pollution is generally considered to result from anthropogenic activities, such as mining, mineral fertilizers, vehicle exhaust, and other industrial activities [2,3]. Industrial activities are regarded as the principal contributor for heavy metal pollution [4]. With the rapid industrialization and urbanization of China over the last decades, various industrial activities have contributed a large amount of heavy metals to the soil, directly and indirectly [5]. The urban soil around electronics manufacturing has been subject to multiple heavy metal contaminations in the Hebei province of China [6]. Wastewater from unregulated manufacturers



is the primary heavy metal contributor in urban river systems [7]. It has been found that about 50% of the soil samples are contaminated by heavy metals in the gold-mining region of Shanxi province, China [8]. Heavy metal exposure in ecosystems can endanger both animals and plants, and harm human health via the food chain [9–11]. Long-term exposure to heavy metals has been associated with problems such as hearing loss, intellectual disabilities, nervous system dysfunction, behavioral problems, and various cancers. Furthermore, exposure to multiple heavy metals may induce more severe diseases such as immune system damage, skin cancer, skeletal damage, vascular disease, and so on [12,13].

Over the past decade, in order to reduce the heavy metal risks to human health, the remediation of heavy metal contaminated soils has been a worldwide environmental goal. Various technologies (surfication, landfilling, soil flushing, electro kinetic, extraction, phytoremediation, bioremediation, have been applied to remediate heavy metal contaminated soils. However, most of these techniques are costly and may cause secondary pollution [14]. Phytoremediation is a promising method for the removal of heavy metals from contaminated soils, which is considered a low-cost, effective, and environmental friendly approach for remediation, and has been widely adopted over the world [15]. In general, phytoremediation is classified into several subcategories, namely: phytostabilization, phytoextraction, phytovolatilization, phytodegradation, and so on. Phytostabilization minimizes or restricts the movement of pollutants by plants, phytoextraction relies on plants to absorb soil heavy metals, phytovolatilization removes volatile pollutants or metabolites using plants, and phytodegradation degrades organic pollutants by inducing the metabolic activities of plant root microbes [16]. Although in the future, the genetic engineering of plants may help improving phytoremediation [17], presently, there is an urgent demand to select promising species from native plants [18–20]. The use of native plants is a valuable option, because these plants are better adapted to the regional multi-stressful environment than the introduced ones [21,22]. Previous studies also reported that heavy metal accumulators were usually found in metal-contaminated environments [23,24]. Therefore, it could be an effective way to assess the phytoremediation potential of native plants in a metal-contaminated environment.

With the implementation of China's Western Development Policy, more industrial areas have improved their economies in the northwest. However, the soil quality has been severely deteriorated, especially because of the heavy metal soil pollution. Most industrial parks have been built on the edges of rivers to have better access to water resources, thus causing severe heavy metal pollution in rivers, because of surface runoff and the mobility of heavy metal in soils [25]. As one of the fastest-growing cities with an industrial economy, Shizhuishan, located in northwestern Ningxia, China, is considered as a typical city that has exhausted coal as a resource. The leading industries include mining, smelting, electroplating, energy, chemical, fuel production, and power transmission. Despite the economic development generated from these industries, it is well known that these industries can lead to serious heavy metal contamination by discharging waste residue into soils and waste water into the river. Wang et al. [26] found that the soil heavy metal pollution in the industrial area of Ningxia is the most serious in different functional zones. A previous study by Zhou et al. [27] showed that the groundwater heavy metal pollution in Ningxia has been worsened by industry. Heavy metals dissolved in water are easily absorbed by organisms and can be bio-accumulated into the food chain. The long-term exposure to heavy metals for humans may affect growth, metabolism, reproduction, and even lead to various diseases. Therefore, it is urgent to find and evaluate the heavy metal pollution of surface soils in the Shizuishan industrial district. It is also necessary to further effectively remediate the heavy metal pollution by phytoremediation technology. Phytoremediation utilizes plants to clean the heavy metal contamination. Many plants have been reported to tolerate and accumulate heavy metals, and can be used to eliminate the heavy metal contamination in soils [28,29]. However, plants that grow in the arid zone of northwest China are subjected not only to heavy metals, but to saline conditions and sometimes drought 30. Heavy metal accumulators from introduced ones cannot survive in these multiple environmental stresses. Thus, it is the best option to use native plants for phytoremediation,

as they grow well in the harsh environment. The main objectives of this study were (1) to assess heavy metal pollution through different methods and (2) to identify potential phytoremediator of Cd, Cr, As, Pb, and Cu for phytoremediation in this region.

2. Materials and Methods

2.1. Study Area

Shizuishan is a typical industrial city in the northern part of Ningxia Province, China, which is bordered by the Yellow River in the east and Henlan Mountain in the west. There are three large industrial parks in this district, including Hebin industrial park (a), Hongguozi industrial park (b), and the Agricultural processing industrial park (c) (Figure 1); heavy metals mainly originated from the factories of the three industrial parks in the region. The area is characterized by strong solar radiation, frequent wind, and dry air. Climatic regime is a typical temperate continental monsoon climate, with an annual precipitation of 167.5–188.8 mm and a potential evaporation of 1708.7–2512.6 mm. The main soil type is gray desert soil [30]. The soil tends to be sandy textured and have pH values ranging from 8.0 to 9.1. The landscape is dominated by desert grassland, and the vegetation coverage fraction is very low in the study area. Herbaceous plants are the dominant vegetation type, such as *Artemisia verbenacea, Peganum harmala, Salsola collina Pall*, and so on.

2.2. Sample Collection and Analysis

To obtain representative data for the three industrial parks, 45 soil samples were collected from two sample zones (A and B) in August 2017. Based on the location of the three industrial parks, we designed zone A between the Hebin industrial park and the left bank of the Yellow River, and zone B between the center of Hongguozi and the agricultural product processing industrial park to the left bank of the Yellow River. There were 12 and 33 samples that were collected in zone A and B, respectively, by line transects of systematic sampling, with 800 m apart between the two sample locations (Figure 1).

Each sample was identified using a global positioning system (GPS) device to determine its longitude and latitude. The surface soil samples were collected between depths of 0 to 10 cm, and each composited sample (approximately 500 g) consisted of soils collected at the central point and four additional points within the radius of 2.5 m towards the north, east, south, and west.

At every soil sample point, plant species were investigated within an area of 1 m². The height, coverage, and number of plants encountered were recorded and their important value was calculated. The important value is an indicator of the species roles in community. It is the sum of the relative height, relative frequency, and relative coverage of a species [31]. According to the important value, 21 species were collected from the corresponding soil sample locations, and three to five individuals of each species were randomly selected from the sample points. In the study area, herb species were the dominant life type, while there were rarely shrubs. There were 38 plant species of 11 families that were recorded, mainly composed of *Asteraceae*, *Gramineae*, and *Chenopodiaceae*, which accounted for 26.3%, 21.1%, and 21.1% of the total species, respectively (Table 1). These plants have common characteristics is dust removal, contamination resistance to heavy metals, and adaptability [32]. These plants were collected and stored in a cooler and transported to the lab immediately.

The concentration of Cd, Cr, As, Pb, Cu, and Zn in soils and plants were determined in the Key Laboratory of Eco-environment in the Three Gorges Reservoir Region of the Ministry of Education, College of Life Sciences, Southwest University, Chongqing, China. The soil samples were air-dried at room temperature, grinded and passed through a 100-mesh plastic sieve, and then oven-dried at 70 °C for 24 h. The fresh plants were separated into roots and shoots, carefully washed with deionized water, oven-dried at 70 °C for 24 h, and then the dry weight of every plant was measured after grinding into fine powder using a ball mill. For the analyses of the six heavy metals in the plant roots and shoots, approximately 0.05 g of material was digested by a microwave with a mixture of HNO_3/H_2O_2

(3:1). Similarly, 0.05 g of soil was digested using a mixture of $HNO_3/H_2O_2/HF$ (7:2:1). The total concentrations of Cd, Cr, As, Pb, Cu, and Zn were determined by inductively coupled plasma-optical emission spectroscopy (ICP-OES, Thermo Fisher iCAP 6300, Loughborough, UK) [33].



Figure 1. Distribution of the sample locations in the study area. The left panel shows Shizuishan city of Ningxia, China; the top right panel shows sample zone A and B of Huinong district in Shizuishan city: (a) Hebin industrial park, (b) Hongguozi industrial park, and the (c) agricultural product processing industrial park. The last panel shows the sample distribution in zones A and B.

Family	Species	Important Value	Classification
Gramineae	Agropyron cristatum	0.040	Perennial
	Achnatherum splendens	0.034	Perennial
	Chloris virgata	0.060	Annual herb
	Echinochloa crusgalli	0.110	Annual herb
	Leymus secalinus	0.043	Perennial
	Phragmites japonica	0.101	Perennial
	Setaria viridis	0.071	Annual herb
	Tragus racemosus	0.026	Annual herb
Zygophyllaceae	Peganum harmala	0.090	Perennial
	Tribulus terrester	0.064	Annual herb
Asteraceae	Artemisia blepharolepis	0.084	Perennial
	Artemisia scoparia	0.087	Perennial
	Artemisia verbenacea	0.097	Perennial
	Cirsium setosum	0.036	Perennial

Table 1. The species characteristics of the study areas.

Family	Species	Important Value	Classification
	Mulgedium tataricum	0.021	Perennial
	Scorzonera divaricata	0.036	Perennial
	Sonchus oleraceus	0.038	Annual herb
	Xanthium sibiricum	0.053	Annual herb
Chenopodiaceae	Bassia dasyphylla	0.053	Annual herb
	Chenopodium album	0.048	Annual herb
	Chenopodium glaucum	0.067	Annual herb
	Chenopodium serotinum	0.052	Annual herb
	Halogeton arachnoideus	0.080	Annual herb
	Kochia scoparia	0.050	Annual herb
	Salsola collina	0.093	Annual herb
	Salicornia europaea	0.028	Annual herb
	Suaeda glauca	0.060	Annual herb
	Suaeda salsa	0.067	Annual herb
Polygonaceae	Polygonum aviculare	0.045	Annual herb
Amaranthaceae Amaranthus retroflexus		0.037	Annual herb
Portulacaceae	Portulaca oleracea	0.020	Annual herb
Convolvulaceae	Convolvulus arvensis	0.054	Perennial
Typhaceae	Typha orientalis	0.038	Perennial
Asclepiadaceae	Cynanchum chinense	0.057	Perennial
Leguminosae	Caragana stenophylla	0.028	Shrub
	Medicago sativa.	0.034	Perennial
	Lespedeza bicolor	0.041	Shrub
	Glycyrrhiza uralensis	0.036	Perennial

Table 1. Cont.

2.3. Assessment of Soil Pollution

As a result of the alkaline nature of the soil in the study region, the six heavy metals were assessed using the second level standards of the Environmental Quality Standard for Soils (GBl5618-1995). The second level standards of Cd, Cr, As, Pb, Cu, and Zn were 0.6, 250, 25, 350, 100, and 300 mg·kg⁻¹, respectively [34].

The single factor pollution index [35] was used to assess the pollution level of a single heavy metal. The Nemerow synthetic pollution index [36] was used to assess the overall pollution caused by the simultaneous presence of several heavy metals, which incorporates the mean and the maximum value of a single factor pollution index. Different heavy metal pollutions have different impacts on the environment, thus the weight coefficient of different heavy metals must be considered. This study adopted the weight coefficient suggested by Swaine [37]. To be specific, Cd, As, and Pb fell into the first category, which were the greatest environmental threats, and had a weight coefficient of three, whereas Cr, Cu, and Zn were in the second category with a weight coefficients of two.

(1) The single factor pollution index is expressed as follows:

$$P_i = C_i / S_i$$

where P_i is the single factor pollution index of heavy metal *i*, and a larger P_i value indicates that the heavy metal pollution of soils is more serious. C_i is the measured value of the heavy metal *i*. S_i is the second level standard of the Environmental Quality Standard for soils of heavy metal *i*.

(2) The Nemerow synthetic pollution index is expressed as follows:

$$P_N = \sqrt{\frac{(C_i/S_i)_{\max}^2 + (\sum_{i=1}^n w_i P_i / \sum_{i=1}^n w_i)^2}{2}}$$

where P_N is the Nemerow synthetic pollution index in every sampling site, $(C_i / S_i)_{\text{max}}$ is the corresponding maximum value in the single factor pollution index, $\sum_{i=1}^{n} w_i P_i / \sum_{i=1}^{n} w_i$ is the corresponding weighted average value in the single factor pollution index, and w_i is the weight coefficient of different heavy metals. The grade standard of the single factor pollution index and the Nemerow synthetic pollution index are showed in Table 2 [35,36].

Single Factor Nemerow Pollution Pollution Grade Pollution Grade Grade Index (P_i) Index (P_N) $P_N \le 0.7$ $P_i \leq 1$ No pollution 1 Clean $0.7 < P_N \le 1$ 2 $1 < P_i \leq 2$ Low pollution Warn limit 3 $2 < P_i \leq 3$ $1 < P_N \leq 2$ Moderate pollution Slight pollution 4 $2 < P_N \leq 3$ $P_i > 3$ High pollution Moderate pollution 5 $P_N > 3$ Heavy pollution

Table 2. The grade standard for soil heavy metal pollution.

Note: P_i is the single factor pollution index of heavy metal *i* and P_N is the Nemerow synthetic pollution index in every sampling site.

The bioconcentration factor (BCF) and translocation factor (TF) are useful evaluate whether a particular plant is a heavy metal hyperaccumulator [38]. The ability of a plant to accumulate heavy metals from soils can be estimated using the BCF, and the ability of a plant to transfer metal from the root to the shoot is measured using the TF. BCF and TF [39,40] were calculated as follows:

$$BCF = C_P / C_S \times 100\%$$

where C_p is the heavy metal concentration in the whole plant and C_s is the heavy metal concentration in the soil (mg·kg⁻¹ DW).

$$\mathrm{TF} = C_s / C_r \times 100\%$$

where C_s is the heavy metal concentration in the shoot of a plant and C_r is the heavy metal concentration in the root (mg·kg⁻¹ DW).

2.4. Statistical Analysis

Statistical analyses were performed for all of the data using SPSS 20.0 and Excel 2010, and figures were drawn using Origin 8.5. The Pearson correlation analyses were performed to establish the relationships of the heavy metals between the soil and plants. Differences between the different plants on the enrichment capability and transfer ability of heavy metals were studied with one-way analysis of variance (ANOVA) with Duncan's multiple range test at the 5% level. All of the datasets were normal in our study.

3. Results

3.1. Heavy Metal Concentration in Soils

The descriptive statistics of six heavy metal contents in soils arre presented in Table 3.

Elements	Range	Mean \pm SE	Coefficient of Variation (%)	Distribution Type	Soil Background Content in Ningxia	Over Standard Rate ⁽¹⁾ (%)
Cd	2.1~8.5	4.06 ± 0.20	33.61	normal	0.11	100.00
Cr	59.8~132.3	100.27 ± 1.92	12.86	normal	60.00	97.78
As	60.3~145.1	91.40 ± 3.08	22.62	normal	12.70	100.00
Pb	18.2~81.6	28.50 ± 1.62	38.03	normal	20.60	88.89
Cu	18.9~42.4	28.16 ± 0.695	16.56	normal	22.10	91.11
Zn	222.6~664.2	391.37 ± 16.08	27.56	normal	58.80	100.00

Table 3. Characteristics of soil heavy metals in study areas (n = 45; mg·kg⁻¹).

Note: ⁽¹⁾ The standard is soil background values in Ningxia.

As a whole, the mean value of the six heavy metal contents in the soils followed a descending order of Zn > Cr > As > Pb > Cu > Cd. All of the metal concentrations were far higher than their background values in Ningxia. Relatively, they were 36.91, 1.67, 7.20, 1.38, 1.27, and 6.66 times that of the corresponding background values, respectively. Based on their background values [41], the overall standard rate of six heavy metals were 90% higher, while Cd, As, and Zn were 100% higher. The coefficients of variation varied from 12.86% for Cr to 38.03% for Pb, and decreased in the order of Pb > Cd > Zn > As > Cu > Cr (Table 3).

3.2. Pollution Assessment of Heavy Metals

The single factor pollution index and Nemerow synthetic pollution index for the six heavy metals that were measured are summarized in Table 4.

	P_i						л
	Cd	Cr	As	Pb	Cu	Zn	P_N
Max	14.1	0.53	5.8	0.23	0.42	2.21	10.47
Min	3.5	0.24	2.4	0.05	0.19	0.74	2.67
Mean	6.77	0.40	3.66	0.08	0.28	1.30	5.07
Pollution level	Heavy pollution	Unpolluted	Heavy pollution	Unpolluted	Unpolluted	Light pollution	Serious pollution

Table 4. Single factor index (P_i) and Nemerow pollution index (P_N) for heavy metals.

Note: P_i is the single factor pollution index of heavy metal *I* and P_N is the Nemerow synthetic pollution index in every sampling site.

The average single factor pollution index for the six heavy metals decreased in the order of Cd > As > Zn > Cr > Cu > Pb. The average single factor pollution index for Cd and As were greater than three, showing severe pollution. The values of Cd in all of the sampling sites ranged from 3.5 to 14.1, indicating serious contamination at all sites. The average single factor pollution index for Zn was between one and two, indicating that the study areas were slightly polluted by Zn. The maximum single factor pollution indices of Cr, Pb, and Cu in all of the samples were 0.40, 0.08, and 0.28, respectively. These values were lower than one, indicating that all of the samples were not polluted by Cr, Pb, and Cu. The average Nemerow synthetic pollution index in the industrial area was higher than three, which was serious pollution. On the whole, our data show that 97.78% of the samples were seriously polluted, and the rest were moderately polluted.

3.3. Relationship between Metal Levels in Soil and Plants

The correlation coefficients of the six heavy metals between the soils and plants are presented in Table 5. For *Artemisia blepharolepis*, *Suaeda salsa*, *Mulgedium tataricum*, *Leymus secalinus*, and *Chloris virgata*, the Cd content between the soils and plants showed a higher significant positive correlation. *Polygonum aviculare*, *Amaranthus retroflexus*, *Chenopodium glaucum*, *Tribulus terrester*, *Chenopodium album*, and *Achnatherum splendens*, their Cr content showed a higher significant positive correlation with corresponding soil, Cr. For *Halogeton arachnoideus*, *Polygonum aviculare*, *Bassia dasyphylla*, *Suaeda salsa*, *Salsola collina*, *Mulgedium tataricum*, and *Chloris virgata*, the Pb content between the soils and plants showed a higher significant positive correlation. For *Halogeton arachnoideus*, *Echinochloa crusgalli*, *Amaranthus retroflexus*, *Setaria viridis*, *Artemisia scoparia*, *Achnatherum splendens*, and *Chloris virgata*, the Cu content between the soils and plants showed a higher significant positive correlation. For *Halogeton arachnoideus*, *Polygonum aviculare*, *Tribulus terrester*, and *Kochia scoparia*, the Zn content between the soils and plants showed a higher significant positive correlation at the 0.05 probability level.

Species			r		
	Cd	Cr	Pb	Cu	Zn
Artemisia blepharolepis	0.821 *	-0.562	-0.998 *	-0.996 *	0.421
Amaranthus retroflexus	0.027	0.727 *	-0.832 *	0.759 *	-0.669
Artemisia scoparia	0.295	0.181	-0.491	0.523	0.454
Achnatherum splendens	-0.675	0.573 *	-0.825 *	0.987 *	0.029
Bassia dasyphylla	0.324	-0.064	0.570	0.318	0.063
Chenopodium album	-0.843 *	0.960 *	0.288	-0.739	-0.506
Chenopodium glaucum	-0.370	0.776 *	0.129	-0.180	-0.216
Chloris virgata	0.915 *	-0.730	0.911 *	0.991 *	-0.736 *
Echinochloa crusgalli	-0.642 *	-0.282	0.383	0.823 *	0.350
Halogeton arachnoideus	0.010	-0.418	0.938 *	0.883 *	0.966 *
Kochia scoparia	-0.052	-0.857 *	-0.632	-0.234	0.680 *
Leymus secalinus	0.901 *	-0.811	-0.762	-0.766 *	-0.663
Mulgedium tataricum	0.984 *	-0.840 *	0.979 *	-0.723	-0.251
Polygonum aviculare	-0.804 *	0.901 *	0.998 *	0.329	0.658 *
Peganum harmala	0.173	-0.668	-0.246	0.072	0.027
Phragmites japonica	0.141	-0.118	-0.333	0.087	-0.393
Salsola collina	0.297	0.411	0.884 *	0.466	0.108
Suaeda glauca	0.050	-0.691	-0.342	-0.101	-0.578
Suaeda salsa	0.547	0.225	0.559 *	-0.250	-0.698 *
Setaria viridis	-0.839 *	-0.818 *	-0.945 *	0.886 *	-0.608
Tribulus terrester	0.007	0.857 *	0.424	-0.329	0.716 *

Table 5. The correlation analysis between metal levels in soil and plants.

Note: r showed that the correlation coefficients between the metal levels in the soils and plants. * showed that the correlation coefficients had statistical significance at the 0.05 probability level (p < 0.05).

3.4. Heavy Metal Concentration in Plants

The concentration of heavy metals in the different plant samples are given in Table 6. For metal Cd, *Artemisia blepharolepis* and *Leymus secalinus* presented a higher accumulation, due to their higher biomass, than other plants. However, the Cd concentrations in the shoot and root of all of the studied plants were limited. The Cr concentrations in the shoots among the studied species had no significant differences. But *Achnatherum splendens* presented a higher level of Cr in the root compared with the other plants (p > 0.05). *Chloris virgata* had a higher value of metal Pb in the root, but the low biomass limited the application in phytoremediation. *Halogeton arachnoideus* and *Salsola collina* showed a relatively higher biomass and Pb contents than the other plants. For Cu, the concentrations in the roots among the studied species showed no significant differences, and *Artemisia scoparia* and *Echinochloa crusgalli* had significantly higher concentrations in the shoots (p > 0.05). *Tribulus terrester* and *Halogeton arachnoideus* showed higher levels of metal Zn in shoots and roots than other plants, but the biomass of *Tribulus terrester* was lower than that of *Halogeton arachnoideus*.

Elemente	Species	$\mathbf{P}_{ionum} = \mathbf{r}_{ionum} = \mathbf{r}_{ionum}$	Heavy Metals Concentration(mg⋅kg ⁻¹)		
Elements		biomass(g·plant -)-	Shoot	Root	
Cd	Artemisia blepharolepis	46.10	$0.45\pm0.09~\mathrm{a}$	0.60 ± 0.20 a	
	Chloris virgata	2.49	$0.19\pm0.01~\mathrm{a}$	$0.50\pm0.06~\mathrm{a}$	
	Leymus secalinus	23.75	$0.91\pm0.12~\mathrm{a}$	$0.70\pm0.06~\mathrm{a}$	
	Mulgedium tataricum	5.81	$0.29\pm0.02~\mathrm{a}$	$0.30\pm0.03~\mathrm{a}$	
	Suaeda salsa	16.48	$0.75\pm0.29~\mathrm{a}$	$0.45\pm0.07~\mathrm{a}$	
Cr	Achnatherum splendens	38.10	$55.21\pm3.85~\mathrm{a}$	$152.35\pm8.60~\mathrm{b}$	
	Amaranthus retroflexus	10.80	$34.70\pm7.14~\mathrm{a}$	$21.24\pm4.33~\mathrm{a}$	
	Chenopodium album	14.30	$22.41\pm2.20~\mathrm{a}$	$69.28\pm10.20~\mathrm{a}$	
	Chenopodium glaucum	11.28	60.74 ± 22.09 a	59.46 ± 15.65 a	
	Polygonum aviculare	13.73	$63.37\pm46.86~\mathrm{a}$	44.92 ± 23.47 a	
	Tribulus terrester	6.85	$62.27\pm18.26~\mathrm{a}$	66.26 ± 31.13 a	
Pb	Bassia dasyphylla	7.40	$5.93\pm1.47~\mathrm{a}$	$2.30\pm0.49~\mathrm{a}$	
	Chloris virgata	2.49	2.46 ± 0.14 a	$10.78\pm0.08~\mathrm{c}$	
	Halogeton arachnoideus	16.80	$5.86\pm0.59~\mathrm{a}$	$5.32\pm0.27\mathrm{b}$	
	Mulgedium tataricum	5.81	$3.43\pm0.32~\mathrm{a}$	$2.98\pm0.07~\mathrm{ab}$	
	Polygonum aviculare	13.73	4.68 ± 0.34 a	$3.93\pm0.57~\mathrm{ab}$	
	Salsola collina	16.20	$4.86\pm0.78~\mathrm{a}$	$5.98\pm0.91\mathrm{b}$	
	Suaeda salsa	16.48	2.56 ± 0.34 a	$2.37\pm0.26~\mathrm{ab}$	
Cu	Achnatherum splendens	38.10	$5.72\pm0.25~\mathrm{a}$	$9.82\pm0.22~\mathrm{a}$	
	Amaranthus retroflexus	10.80	$8.31\pm1.18~\mathrm{ab}$	$10.01\pm3.14~\mathrm{a}$	
	Artemisia scoparia	16.80	$14.00\pm3.27~\mathrm{c}$	11.27 ± 0.70 a	
	Chloris virgata	2.49	$5.78\pm0.59~\mathrm{a}$	10.70 ± 0.61 a	
	Echinochloa crusgalli	16.07	$12.44\pm0.80~bc$	20.55 ± 3.55 a	
	Halogeton arachnoideus	16.80	$6.87\pm0.76~\mathrm{ab}$	8.47 ± 0.88 a	
	Setaria viridis	25.89	$6.38\pm1.73~\mathrm{ab}$	$15.67\pm1.17~\mathrm{a}$	
Zn	Halogeton arachnoideus	16.80	$91.53\pm26.59~\mathrm{b}$	$64.03 \pm 17.51 \text{ a}$	
	Kochia scoparia	13.60	20.46 ± 0.93 a	$23.23\pm3.31~\mathrm{a}$	
	Polygonum aviculare	13.73	$72.16\pm3.20~\text{ab}$	$60.02\pm14.25~\mathrm{a}$	
	Tribulus terrester	6.85	$93.02\pm8.90b$	$110.71\pm6.91\mathrm{b}$	

Table 6. Heavy metal concentrations in the shoots and roots of different plants.

Note: data of the heavy metal concentration were showed for the means \pm standard error with one-way analysis of variance. Lowercase letters indicated heavy metal concentrations with significant differences at the 0.05 probability level.

3.5. Bioconcentration and Translocation Factors in Native Plants

The ANOVA results showed that the bioconcentration factor and translocation factor were significantly different in different plants; the Duncan's test results are shown in Figure 2 by capital and small letters.



Figure 2. Bioconcentration factor (BCF) and translocation factors (TF) of five heavy metals found in different plants. The data were showed for the means \pm standard error with one-way analysis of variance. Lowercase letters indicated that the BCF of different species has statistically significant differences at the 0.05 probability level. Uppercase letters indicated that the TF of different species has statistically significant differences at the 0.05 probability level. BCF—bioconcentration factor; TF—translocation factor. The red lines showed the threshold line of BCF and TF.

The bioconcentration factors of Cd in the order of Leynus secalinus > Suaeda salsa > Artemisia *blepharolepis* > *Chloris virgata* > *Mulgedium tataricum*, and the top three species, were not significantly different at the 0.05 probability level. The translocation factors of Suaeda salsa and Leymus secalinus had the same significance (p < 0.05), but *Suaeda salsa* was the highest. In the studied plants for Cr, the bioconcentration factors were lower than one. Achnatherum splendens, Tribulus terrester, and *Chenopodium glaucum* were relatively higher than the other plants, and they were not significantly different at the 0.05 probability level. The translocation factors of Cr were in the order of Amaranthus retroflexus > Tribulus terrester > Polygonum aviculare > Chenopodium glaucum > Achnatherum splendens > *Chenopodium album*, and the top four species were not significant at the 0.05 probability level. For Pb, the bioconcentration factors of all of the samples were far lower than one, and they were relatively stable, ranging from 0 to 0.1. These values were not significant at the 0.05 probability level. The translocation factor of Bassia dasyphylla was significantly higher than other plants. The bioconcentration factors of Cu were lower than one in the studied species. The values of Artemisia scoparia and Echinochloa crusgalli were relatively higher than other plants, but they were not significant at the 0.05 probability level. The translocation factor of Artemisia scoparia was significantly higher than the other plants. For Zn, the bioconcentration factor of Tribulus terrester was the highest, closely followed by Halogeton arachnoideus, however, they were not significantly different at the 0.05 probability level. The translocation factors for all of the samples of Zn were not significantly different, and the highest was Halogeton arachnoideus.

4. Discussion

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Most studies used soil samples to monitor the environmental metal levels [42]. This study found that the mean values of Cd, Cr, As, Pb, Cu, and Zn in soils were 36.91, 1.67, 7.20, 1.38, 1.27, and 6.66 times that of the corresponding background values, respectively. Based on their background values, the overall standard rate of the six heavy metals were higher than 90%, of which Cd, As, and Zn were 100% (Table 3). These results indicate that all of the six heavy metal contents were relatively high in the study area and almost all of the study area was contaminated by exogenous pollutants. The CV values were used for the description of global variability [43]. Large CV values indicate a considerable spatial variation and imply a significant input from external sources [44]. A low CV suggests that the nonpoint input is predominant [45]. The coefficients of variation varied from 12.86% for Cr to 38.03% for Pb, and decreased in the order of Pb > Cd > Zn > As > Cu > Cr (Table 3), which suggested that there was a moderate degree of spatial variability of the six heavy metals, and the anthropogenic factors significantly influenced the distribution of Pb and Cd.

The heavy metal of the soil pollution status was evaluated by a single factor pollution index and the Nemerow synthetic pollution index (Table 4). From the single factor pollution index, the Cd and As were found to be the most serious pollutants in the industrial area, especially Cd, as all of the samples were seriously contaminated. The study area was slightly polluted by Zn. Areas were relatively clean of Cr, Pb, and Cu, and all of the samples were not polluted by Cr, Pb, and Cu. As a result of the complexity of the soil, the Nemerow synthetic pollution index was employed to evaluate the comprehensive impact caused by the six heavy metals in soil, rather than a single factor pollution index, which can only reveal the pollution level of one metal [46]. According to the Nemerow synthetic pollution index, the industrial area was seriously polluted. As for the pollution level in every sampling site, the results showed that almost all of the samples in the industrial area were seriously polluted by anthropogenic sources. According to the Nemerow synthetic pollution index model, the high single factor pollution indices of Cd, As, and Zn are the main reasons for heavy metal pollution in this region.

Evaluating the phytoremediation potential of heavy metals in different plants should be utilized in the remediation of heavy metal contaminated soils. In general, it is regarded as a hyperaccumulator when the heavy metal concentrations, BCF, and TF for plants are up to corresponding standards [47–49]. However, these nominal thresholds should not be regarded as the absolute cut-off when the phytoremediation potential is assessed. Plants that grow in the semi-arid and arid regions of northwest China are subjected not only to heavy metal contamination, but also to drought and saline stresses. Thus, many studied heavy metal hyeraccumulators, such as *Thlaspi caerulescens*, *Pteris vittata*, *Reynoutria sachalinensis*, and so on, cannot survive and be used for phytoremediation in this region. Some similar studies for this region were conducted in the laboratory. For example, *Ligustrum obtusifolium* was found to have a high capacity of Pb accumulation and translocation under drought stress [50], and *Buddleja alternifolia* had a great potential application in Cd phytoremediation of arid regions [51]. However, these studies were conducted under controlled single heavy metal stresses, which could not reflect the hash and complex situation in situ. Thus, assessing the potential phytoremediators for heavy metals, their phytoremediation potential is still valuable.

As a whole, the bioconcentration factors of five heavy metals in all of the studied species were lower than one, and presented lower levels in the shoot and root than the accumulator, but they had a stronger tolerance under the regional multi-stressful environment. Some plants with a high biomass would share a high resultant capability for phytoremediation. Based on the comprehensive consideration of heavy metal concentrations, BCF, and TF for native plants, we thought *Achnatherum splendens* for metal Cr presented a phytostabilization potential. It grew very well and was abundant in this study area, and had the highest BCF; what is more, the metal Cr was less distributed in the shoot than in the root. As a result of their higher shoot content and BCF in the metal Cu and Zn than other plants (p < 0.05), and the ability to tolerate the regional multi-stressful environment, *Artemisia scoparia* and *Echinochloa crusgalli* for metal Cu and *Halogeton arachnoideus* for metal Zn could be considered as the most promising species for phytoextraction. Almost all of the selected plants were perennials and had a higher important value, which also contributes to enhanced uptake of metal. Although the TF of some plants in metal Cd and Pb was higher than one, all of the studied species presented a low shoot content and BCF, and had no significant differences. Thus, all of the studied plants were limited as phytoremediators for Cd or Pb contaminated soil. Further research should be done in a wider region.

5. Conclusions

The discharge of heavy metals through various industrial activities is an important cause of soil contamination by heavy metals. As the study area has experienced rapid urbanization and industrialization over the past decades, the problem of heavy metal contamination has also become increasingly prominent. The results suggest that the study area was seriously polluted by Cd and As, slightly polluted by Zn, and was relatively clean for Cr, Pb, and Cu contamination. In addition, almost all of the samples in the study area were seriously polluted. The heavy metal remediation of industrial zones should be an important concern; therefore, strategies should be implemented to ban the discharging or dumping of unqualified industrial waste.

In order to reduce heavy metal risks to the human health of local residents, phytoremediation, a natural, esthetically pleasing, and low-cost technology, has opened a new avenue in the remediation of heavy metal contamination soil. The phytoremediators should adapt to heavy metal contamination, drought, and saline stresses in arid and semi-arid land of northwest China. Our results suggest that because of its low shoot content in metal Cr, BCF was the highest compared with other plants, and its ability to tolerate a regional multi-stressful environment, *Achnatherum splendens* for metal Cr presented phytostabilization potential. And, because of its higher shoot content and BCF, and its ability to tolerate a regional multi-stressful environment, *Artemisia scoparia* and *Echinochloa crusgalli* for metal Cu and *Halogeton arachnoideus* for metal Zn presented a phytoextraction potential. And, as a result of its low shoot and root content and BCF, all of the studied plants were limited as phytoremediators for Cd or Pb contaminated soil.

More importantly, the phytoremediation areas should be fenced off from wildlife to prevent contamination of the food chain. The phytoremediation areas could be combusted as biofuel feedstock after harvesting, and the ashes could be recovered from the metals or concentrated landfilled. In addition, as important characteristic of phytoremediation, time-consuming should be noticed compared with other remediation techniques. More research is needed to obtain fast-growing hyperaccumulators though genetic techniques.

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