

Article

Modeling of Atmospheric Dispersion of Jarosite Particles from Tailing Waste in Mitrovica, Kosovo

Mihone Kerolli Mustafa¹, Jelena Djokic^{1,2,*} and Lidija Ćurković^{3,*} ¹ International Business College Mitrovica, Bislim Bajgora nn, 40000 Mitrovica, Kosovo² Faculty of Technical Sciences, University of Pristina, 7 Princi Millosh, 40000 Mitrovica, Kosovo³ Faculty of Mechanical Engineering and Naval Architecture, University of Zagreb, Ivana Lučića 5, 10000 Zagreb, Croatia

* Correspondence: j.djokic@ibcmirovica.eu (J.D.); lidija.curkovic@fsb.hr (L.Ć.); Tel.: +383-44235479 (J.D.); +385-1-6168-183 (L.Ć.)

Abstract: Most of the zinc producers in the world use the jarosite process to improve zinc recovery and to remove iron as an undesirable constituent of zinc ores. Jarosite waste released from the zinc extraction process in Mitrovica, Kosovo has led to severe environmental problems due to toxic heavy metals. This industrial waste from the Zn hydrometallurgy process was abandoned on an open field, being exposed to meteorological conditions and aging. The chemical composition and grain size distribution of the jarosite waste deposit was determined. Microwave digestion procedures were used on whole jarosite samples for use in inductively coupled plasma optical emission spectrometry trace metal analysis (ICP-OES). In addition, different weathering conditions were considered for testing the emission rate of the particles in the laboratory, including relative humidity, wind speed, and temperature. Terrain properties, urban infrastructure, source formation, and location were used for modeling with the AERMOD View-Gaussian air dispersion model. The modeling results showed a range of pollution exceeding the maximum limits in an area of 3 km in the conditions of southeast wind direction and wind speed exceeding 10 m s⁻¹, heavily polluting the city of Mitrovica.



Citation: Kerolli Mustafa, M.; Djokic, J.; Ćurković, L. Modeling of Atmospheric Dispersion of Jarosite Particles from Tailing Waste in Mitrovica, Kosovo. *Atmosphere* **2022**, *13*, 1690. <https://doi.org/10.3390/atmos13101690>

Academic Editors: Jianmin Chen and Thomas Plocoste

Received: 17 September 2022

Accepted: 12 October 2022

Published: 15 October 2022

Publisher's Note: MDPI stays neutral with regard to jurisdictional claims in published maps and institutional affiliations.



Copyright: © 2022 by the authors. Licensee MDPI, Basel, Switzerland. This article is an open access article distributed under the terms and conditions of the Creative Commons Attribution (CC BY) license (<https://creativecommons.org/licenses/by/4.0/>).

Keywords: jarosite particles; air dispersion model; heavy metals

1. Introduction

Worldwide, the jarosite process is the most widely used technique of iron removal from the acidic process in the zinc industry, where it is often produced as a by-product. The use of jarosite in the zinc industry is followed by the production of iron precipitates, with minimum losses of Zn, Cd and Cu in jarosite precipitates. During the zinc extraction process, a huge amount of jarosite is released as waste, where jarosite is extremely common. Therefore, jarosite is common in diverse environments such as nature (oxidizing zones of sulphide deposits, acid sulfate soil, or clays) and hydrometallurgical waste [1–6]. Even though the function of jarosite as a potential source of acid in mines is not completely understood, the traditional view that acid can be generated in mine waste due to the oxidation of sulphide minerals and dissolution of sulfate salts and sulfate minerals such as jarosite has been brought to attention by several researchers [4–8].

Jarosite waste released from the zinc extraction process may cause serious environmental problems due to the presence of toxic metals, such as Pb, Cd, Zn, and As. The characterization and environmental impact of tailing mining waste have been addressed by various researchers worldwide [4,6,7]; however, fewer researchers have addressed the jarosite waste contribution to atmospheric metal deposition in the area of polluted mining districts [9]. From an X-ray diffractogram of jarosite released from the zinc extraction process, the authors confirmed that the major mineral phase is jarosite (KFe₃(SO₄)₂(OH)₆) and iron sulfate hydrate (2Fe₂O₃SO₃·5H₂O). According to Picazo-Rodríguez et al. [1], the mineral phase in jarosite is hydrophobic and insoluble, while the other phase namely iron

sulfate hydrate is hydrophilic and easily soluble. This characteristic nature makes jarosite hazardous material interesting for further study.

The deposition of jarosite waste in open tailing is not a reliable method because of future environmental costs. Tailings are exposed to weathering, acid mine drainage, and the mobilization of metals. There are many documented cases in which the long-term investigation of the bioavailability of heavy metals in the environment has built up toxic metal concentrations in the soil, water, air, and organisms [1,5,9]. Mining in Kosovo has occurred for a long period, with lead and zinc metallurgy productions in the Trepca mine in Mitrovica in the region. With reserves of lead, zinc, bauxite, nickel, silver, and gold, Trepca is known as one of the biggest mining complexes in Europe [8]. Wind movement over the open tailings in Mitrovica transport a high concentration of heavy metals in the form of particles and present a potential human health concern that requires radical changing measures. Their potential risk is linked with direct exposure to the contaminated site by breathing dust [10–12].

Today, there are several simulation models used for air quality assessment in urban and industrial zones. Atmospheric dispersion models, such as the Gaussian model, Eulerian model, and Lagrangian model, have been applied and implemented using different statistical comparative parameters [10–13]. These models have been used for the assessment of air quality in the mining industry. Most studies have been focused on determining the dispersion of particulate matter but not on assessing the pollution with heavy metals. Castillo et al. reported that mine wastes are the major sources of heavy-metal particles distributed in surrounding areas [11]. The US Environmental Protection Agency proposed several meteorological modeled data sets and models for performing simulations using AERMOD, CALPUFF, and ADMS [14–16]. The AERMOD techniques were used to analyze short- and long-range air dispersion modeling for isolated buildings and wind tunnels and the total suspended particulates from an open-cast coal mine [17]. Zunaira et al. used the quantity of the pollutant's concentration produced at the mining site as the criterion to determine how far the pollutant could travel [17]. The findings showed that Zn in minimal quantities could travel up to 20 km, whereas other heavy metals did not travel beyond 10 km due to lower emissions at the source. Other authors have confirmed that the technique based on the Gaussian algorithm has been used for the mining application [18–20]. The authors have confirmed that the analytical solution and predictions are made by incorporating the emission rate linked to various mining activities [7–20]. It has been observed that there is a need for a more holistic approach to gain more understanding of air pollution multiple-source levels [17,21–24].

The European Environment Agency presented the use of air dispersion models as a response to policymakers and researchers to interpret model results that help the definition of intervention priorities [21]. Evaluation of the modeling results has been the focus of studies in Europe to respond to emission reductions, facilitate air quality planning, and help policymakers in prioritizing policies [20–22]. Very few studies have concentrated on assessing the urban environment exposed to varying concentrations of heavy metals from jarosite [2,5,8]. In this work, we apply the AERMOD View- Gaussian air dispersion model to illustrate to what extent the ratio of the concentration of jarosite particles changes corresponding to the open-tailing location in Mitrovica, Kosovo. The scope of the selected model is to provide the first assessment report on the prediction of air pollutants from jarosite in the form of a model that can be used by the Trepca mining industry and policymakers in Kosovo.

2. Materials and Methods

2.1. Jarosite Waste

Jarosite waste dispersion modeling is a quantitative statistical assessment of the consequences of airborne dust emissions from mine sources in a certain area. Different hydrometeorological conditions, such as wind speed and direction, temperature, relative humidity, and precipitation, are just a few of the variables that affect the transport and

dispersion of jarosite and airborne dust particles. The AERMOD air quality model was used to forecast airborne jarosite waste concentrations. The results are visualized and used for decision-making and analysis purposes.

For that purpose, the following investigations were performed:

- Site characterization, with terrain data, elevation, and locations of the urban infrastructure.
- Processing of terrain data using AERMAP software (PDS Consultancy, Leicester, UK).
- Jarosite deposit characterization as a primary source of emission for environmental impact assessment purposes.
- Processing wind data using AERMET software.
- AERMOD application over an area of interest in order to predict the jarosite particle distribution (plume forecast).

2.2. Site Characterization

The jarosite waste deposit is located on the Sitnica riverbank in the close vicinity of Mitrovica city, Figure 1.

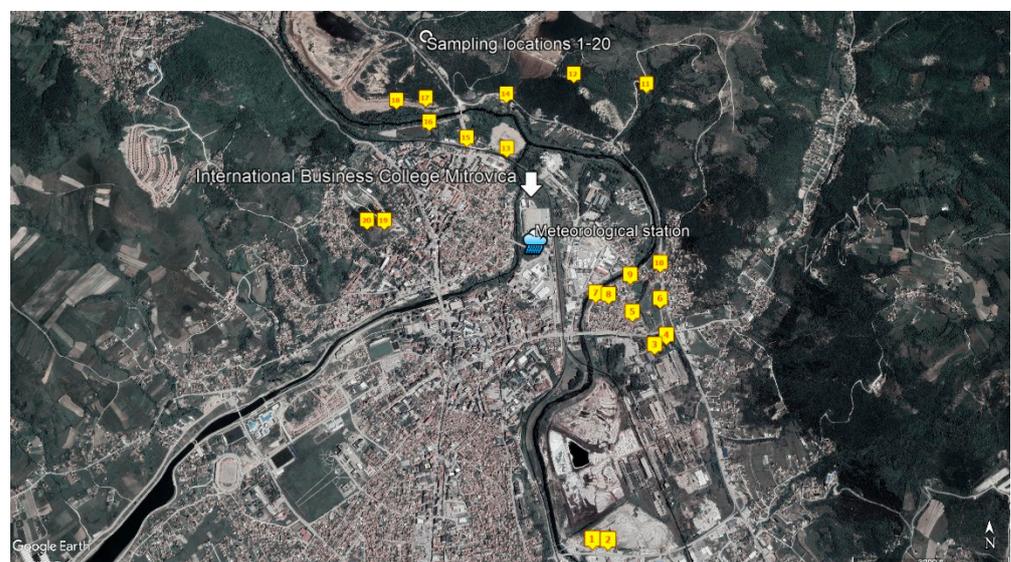


Figure 1. Jarosite waste deposit surrounding terrain with sampling locations (yellow), meteorological station (blue mark), and the location of the probe measuring PM 10 and PM 2.5 at the International Business College building (arrow).

In the Receptor pathway in the AERMOD program, the Flat and Elevated option was selected, as the location is situated at the Sitnica River Valley, which separates the industrial park from the city of Mitrovica, and there are the slopes of the Mountain Bajgora surrounding the site from the eastern side. The AERMOD used the AERMAP Receptor Output file (*ROU). The terrain and location of the broader area of influence, including the location of the nearest meteorological station, are presented in Figure 1. The samples of the topsoil surrounding the jarosite waste deposit were taken from the Sitnica River Valley and analyzed for their heavy metal content, as presented in Table 1.

2.3. Waste Characterization

The jarosite waste deposit is a passive industrial waste deposit, and it was exposed to aging and weathering processes that caused changes in particle size and chemical properties due to agglomeration and oxidation.

The deposit was characterized by chemical analysis and granulometric composition in order to determine the presence of PM 2.5 and PM 10 particles.

Table 1. The amount of the heavy metals (mg kg^{-1}) measured in the top soil of the Sitnica area.

Sample	Pb	Zn	Cd	Cu
	mg kg^{-1}			
1	2245.6	1463	9.98	97
2	6115.7	2061	10.54	259.9
3	40,325	65,510	478.0	325.8
4	1223.1	790	2.47	138.6
5	955.6	1035	5.9	149
6	4969	9866	346.4	230.5
7	5150.5	6328	344.2	253
8	5427.7	4236	229.5	245.2
9	4472.1	6950	124.4	262.5
10	2465.4	567	6.58	166.3
11	2215.5	1252	9.01	95
12	5450	2000	9.23	232.4
13	40,300	64,500	452.0	299.1
14	1115.3	705	1.98	99.98
15	923.4	992	4.30	127
16	4832	9761	332.6	210.3
17	4950	6213	335.5	205
18	5137	3980	211.5	221.2
19	4165.5	6215	105.0	178.5
20	2337.2	459	5.5	135.3
86/278/EEC	300	300	39	140

The total metal content of trace and major metals was determined by means of inductively coupled plasma-optical emission spectrometry ICP=OES (ICP—OES, Teledyne Leeman Labs (Hudson, NH, SAD) after the preparation of samples using the microwave digestion method (MARS 6 XP1500 Microwave Digestion System, CEM, UK). Microwave digestion of jarosite waste samples was carried out in two stages with the use of reagents HCl, HNO₃, HF, and H₃BO₃. Certified reference materials, S JR-3 and S Jsy-1 (MBH Analytical Ltd., Chipping Barnet, UK) were analyzed in order to test the accuracy of the applied method for the determination of total metal concentrations in the investigated jarosite samples. The chemical composition of the jarosite waste deposit is shown in Table 2.

Table 2. Chemical composition of the jarosite waste deposit.

Component	Al ₂ O ₃	Fe ₂ O ₃	SiO ₂	Zn	Pb	Cu	Ag	Ba	Co	Cd	Cr	Mn	Ni	Sr	As
	wt. %			mg kg^{-1}											
amount	1.42	44.94	6.31	10.91	7.51	0.97	133.82	578.9	30.38	2308.7	416.6	6391.7	93.5	156.2	5075.5

The granulometric composition was determined using Zetasizer Ultra (Malvern Panalytical, Malvern, UK) and is shown in Figure 2.

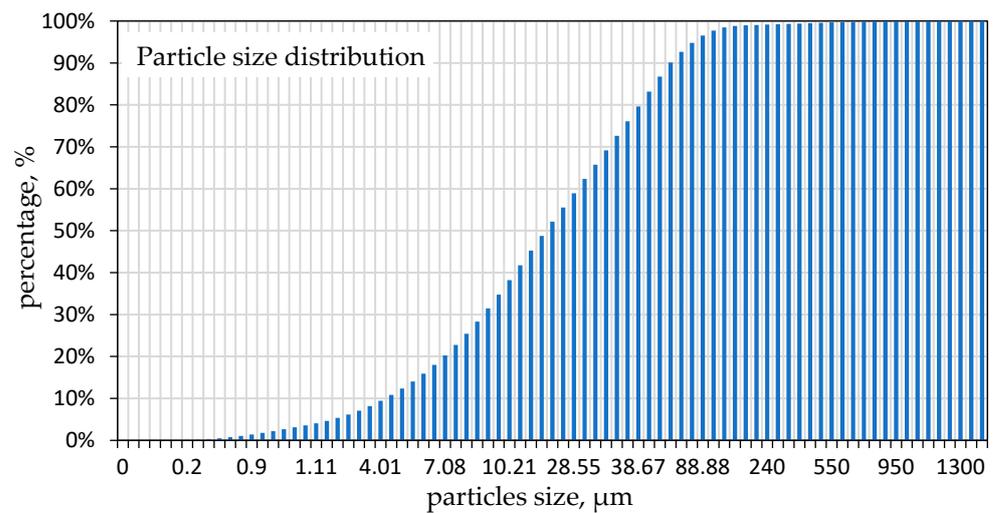


Figure 2. Grain size distribution of the jarosite waste.

As shown in Figure 2, the particle size was smaller than 1 mm, and most of the particles were larger than 10 μm . Less than 8% were PM 2.5, and 30% were PM 10. The rest of them aggregated in weather conditions and aging, so 10% of the particles were larger than 2 mm in diameter. On the other side, the Jarosite waste stored in the open tailing dump became very dry in the summer, which also contributed to airborne dispersion.

For the purpose of modeling, the source was characterized as pollutant TSP and PM 10, Source Type AREA, and polygon was limited by X and Y coordinates.

2.4. Meteorological Data-Climate Characteristics for the Mitrovica Region

In recent years, there has been a trend of decreasing lowest relative humidity data in the summer months. In addition, the humidity data have varied from season to season. The relative humidity data during the last 5 years for the Mitrovica region are presented in Figure 3.

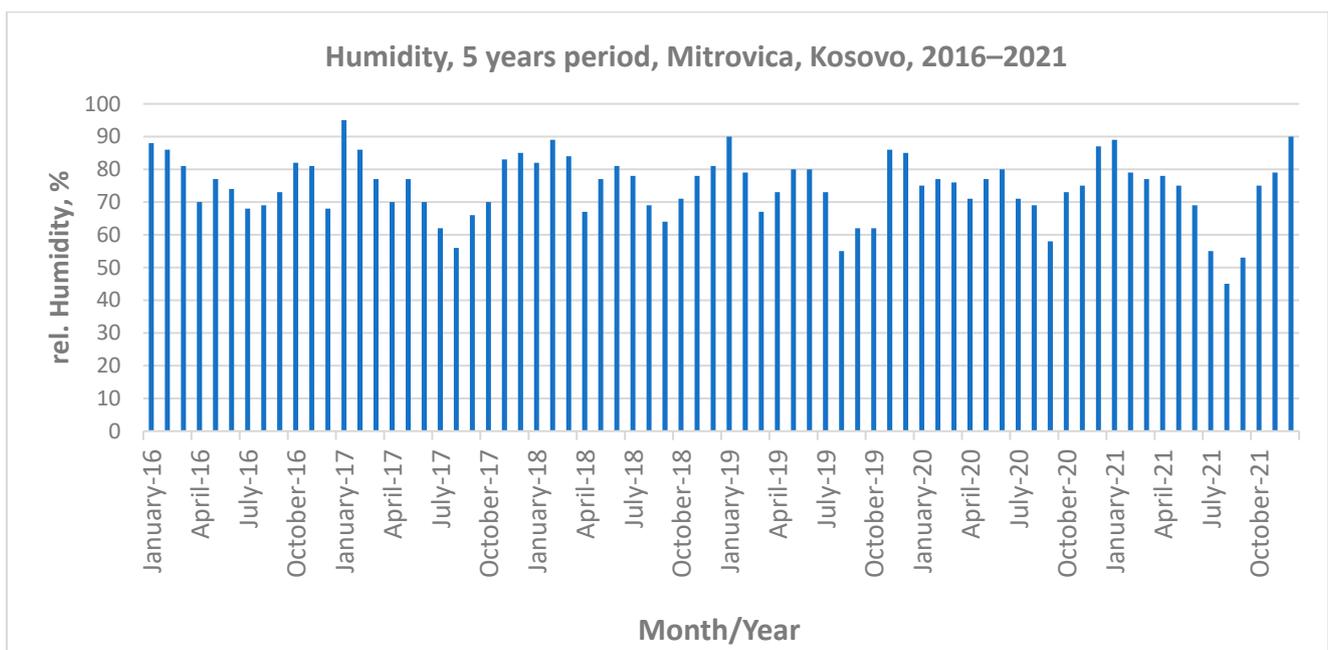


Figure 3. Average relative humidity data in a 5-year period in Mitrovica.

Even though the relative humidity recorded in the last three decades did not show any significant change, there are oscillations in the recorded data in different seasons, showing that winter and autumn seasons bring more precipitation and that July and August are the driest months of the year. The chart shows the average relative humidity for each month during the 5-year period. Overall, 2022 was recorded as extremely dry, the minimum values recorded in August were much lower, and the minimum relative humidity in the close vicinity of the deposit was recorded to be less than 30%, as shown in Figure 4. The data were recorded in the real-time probe in the remote sensing system every 2.5 min, showing significant differences during the day. The results correspond well with the PM 2.5 and PM 10 measurements during August measured at the meteorological station, as shown in Figure 1. For the purpose of including other possible sources of pollution, the average TSP concentrations were taken into consideration. The probe located at the International Business College building collects data, and those data are processed for a 1-year period. The average TSP concentration in the air was calculated to be $21 \mu\text{g m}^{-3}$. Further investigation has determined the origins of these particles as traffic, demolition and construction waste, and individual heating during the winter time.

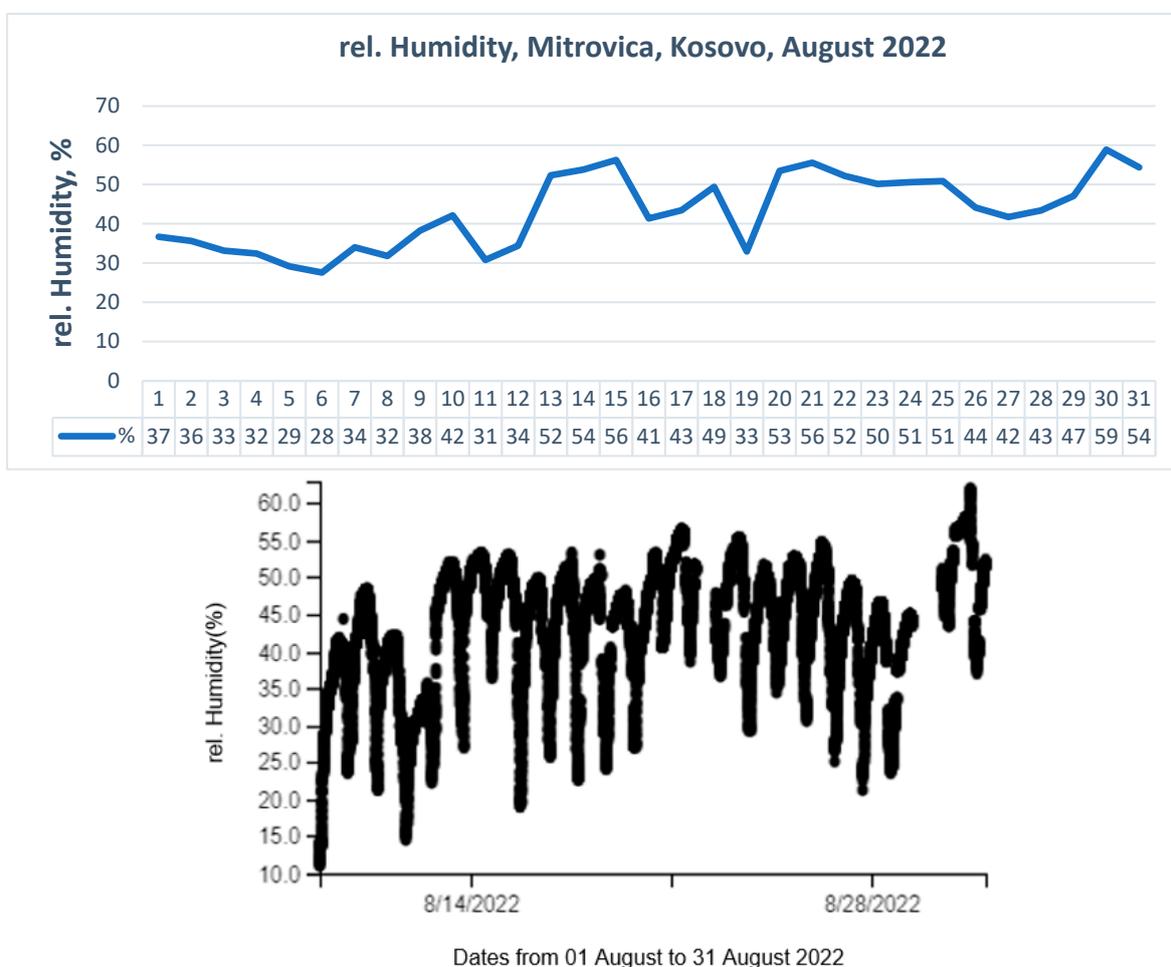


Figure 4. Minimum relative humidity in August 2022.

Meteorological input data were taken from the Meteorological station Mitrovica, Hydrometeorological Institute of Kosovo. The base elevation was 495 m. Figure 5 presents the wind rose for the Mitrovica region based on the wind blowing from the indicated direction. The wind blows from southwest to northeast. Southern winds have a characteristic of the strong west wind, but southwestern winds also increase in frequency and speed toward the

northeast. This situation is reflected during the summer, usually when strong hot winds blow from the Mediterranean area accompanied by hot and dry weather.

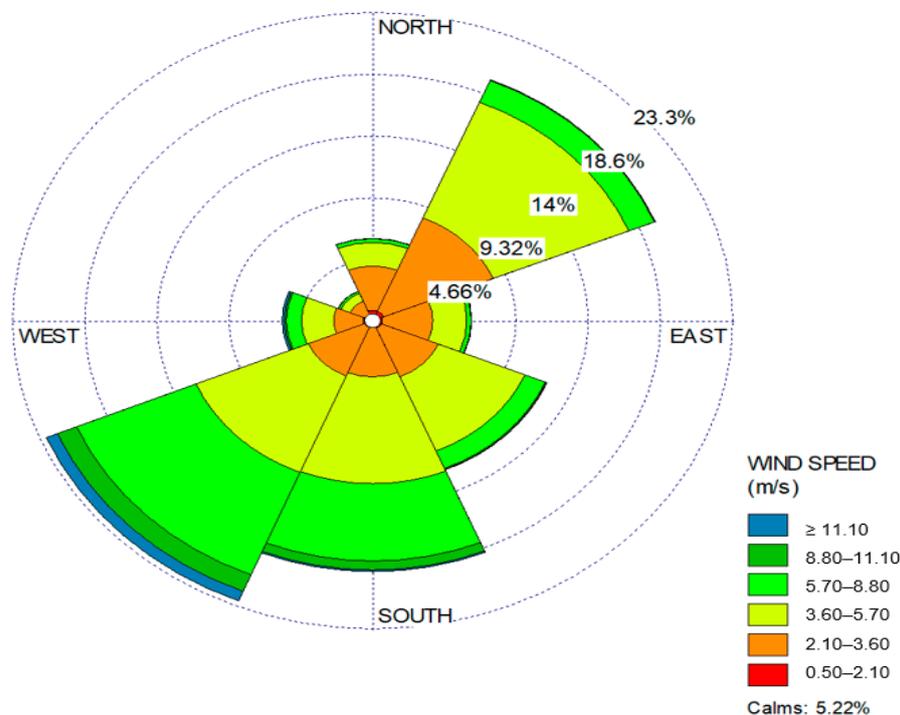


Figure 5. Wind rose plot for the Mitrovica region processed in the WRPLOT supporting program for AERMOD.

2.5. Measurements of Solid Particles Emission within the Given Climate Conditions

The apparatus for the measuring of solid particles carried by the wind was composed of the fan type ABVE-3.5 with a flow of 3600 m³ h⁻¹, a vacuum of 200 Pa used for airflow simulation, and a gravimetric sampler of the respiratory dust. The sample was set in a shallow metal plate with a measuring scale for the captured TSP and digital anemometer (DA-4000). The measurements were performed in the airflow directed from the fan before the apparatus for air vacuuming. The wind speed was changed by changing the distance between the fan and the sample. The experiment was conducted at the wind speed of 5, 7, and 10 m s⁻¹. The humidity in the laboratory was within the interval of 28–56%, and the dusting was 0% (Table 3).

Table 3. Relative humidity and dust emission data.

Relative Humidity	28%	56%
Wind Speed, m s ⁻¹	Dust Emission, g m ⁻² s ⁻¹	
5	0.00013	0.00001
7	0.00021	0.000014
10	0.00035	0.000019

3. Results

In extreme weather conditions, a wind speed stronger than 10 m s⁻¹ is expected, and there are no data on pollution in these climatic conditions. The potentially unstable weather was taken into consideration for modeling. The distance of 5000 m from the source of the pollution jarosite waste deposit was set for discrete distance calculations.

The simulation was performed using the Gaussian Plume Air Dispersion Model AERMOD. AERMOD incorporates, with a new simple approach, current concepts about

flow and dispersion in complex terrain. In this case, the plume is modeled and presents the impacting and following terrain.

Figure 6 shows that there is significant air pollution only in the close vicinity of the deposit. The dashed line marks the jarosite waste deposit. The figure is the result of modeling with the AERMOD software, and it shows the results of the simulation of the particle concentration on the grid of receptors. The receptors grid was set as the Uniform Cartesian grid receptor network, covering an area with a diameter of 6000 m starting from the center of the deposit. The receptors were set at a distance of 200 m from each other, with 961 receptors in total. The Map Projection was set in the Project Coordinate System UTM (Universal Transverse Mercator) with parameters taken from the WGS84 World Geodetic System. Universal Transverse Mercator (UTM) are units used to describe the terrain latitude and longitude, and those units are on the X and Y axes. In Figure 6, the color scale shows the level of concentration, where the concentrations lower than $40,324 \mu\text{g m}^{-3}$ are from yellow to green. The concentrations higher than the EU threshold enter the red part of the spectrum. The constant air dispersion of heavy metals, even in the lower concentration, contributes to soil pollution. The area that was calculated to be the most polluted corresponds with the results of the environmental risk assessment performed with the BCR (the modified European Community Bureau of (BCR) five-step sequential extraction procedure) [5]. The cross-sections were done to establish the relationship between the dispersion and soil pollution in the selected spots. The cross-section in the eastern direction (Figure 7) shows that the air dispersion will be stopped by the hills in the east, and the pollution does not exceed the distance of 800 m in that direction. In addition, the cross-section in the southwestern direction (Figure 8) shows good agreement with the soil analysis presented in Table 1 and shows significant average pollution in the western direction toward the city at a distance larger than 2.4 km. The exact heavy metal content in the particles could be estimated from the total TSP and PM 10 concentration in the air. The grain size distribution of jarosite shows that the particles containing heavy metals participate in the total solid particles with certain percentages: Pb-7.51%; Zn-10.91%; Cd-0.23%; As-0.5%. The average airborne pollution was calculated to be $71 \mu\text{g m}^{-3}$, and the background concentration containing inert dust participation was calculated to be $21 \mu\text{g m}^{-3}$. Therefore, it can be assumed with reasonable certainty that in the remaining $50 \mu\text{g m}^{-3}$, lead will participate with $3.75 \mu\text{g m}^{-3}$, Zinc with $5.45 \mu\text{g m}^{-3}$, Cd with 0.115, and As with $0.25 \mu\text{g m}^{-3}$.

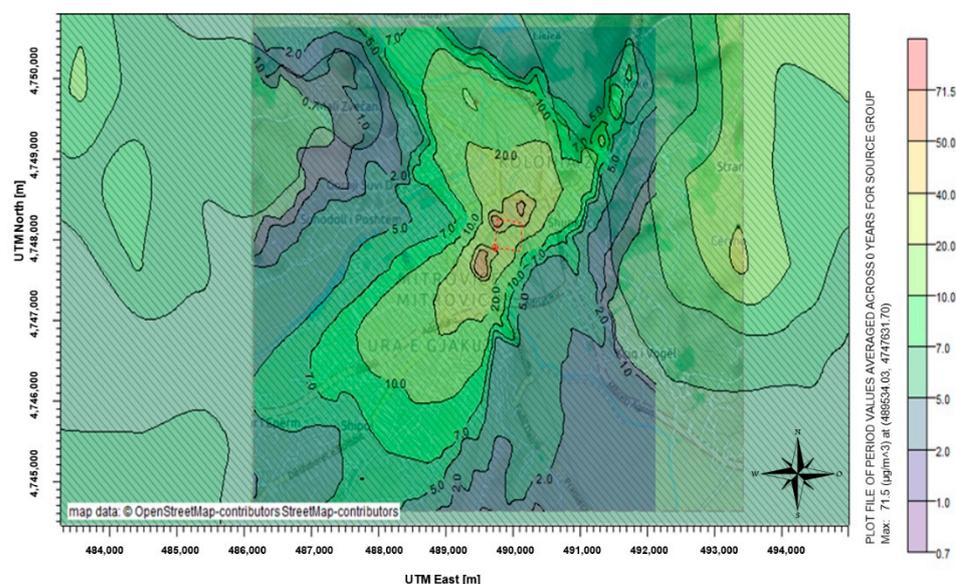


Figure 6. Average jarosite dust concentration.

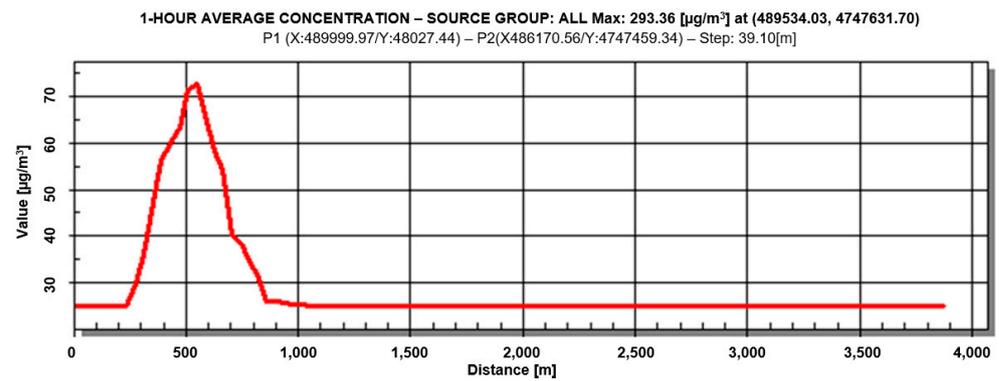


Figure 7. Cross section of the airborne pollution distribution in the east direction.

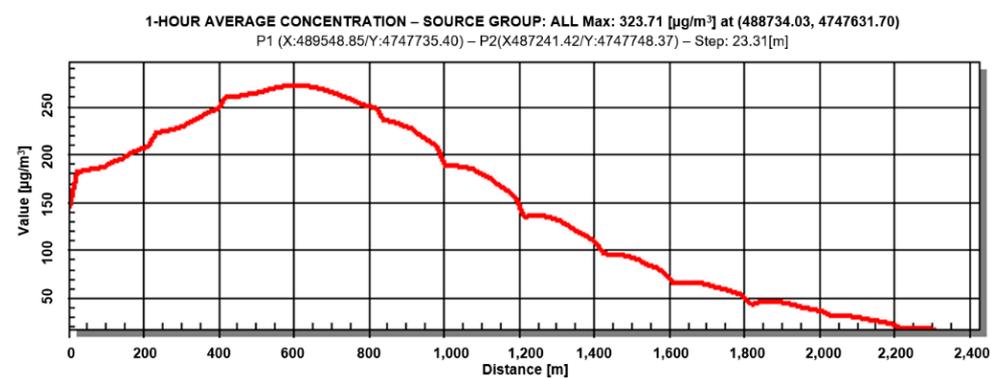


Figure 8. Cross section of the airborne pollution distribution in the southwest direction.

The maximum concentration was calculated with the southwest wind direction and the plot with the PM 10 concentration. Figure 9 shows that the city will become very endangered in the case of the wind blowing from the eastern mountains toward the river valley in the southwest direction, which corresponds well with the map shown in Figure 10.

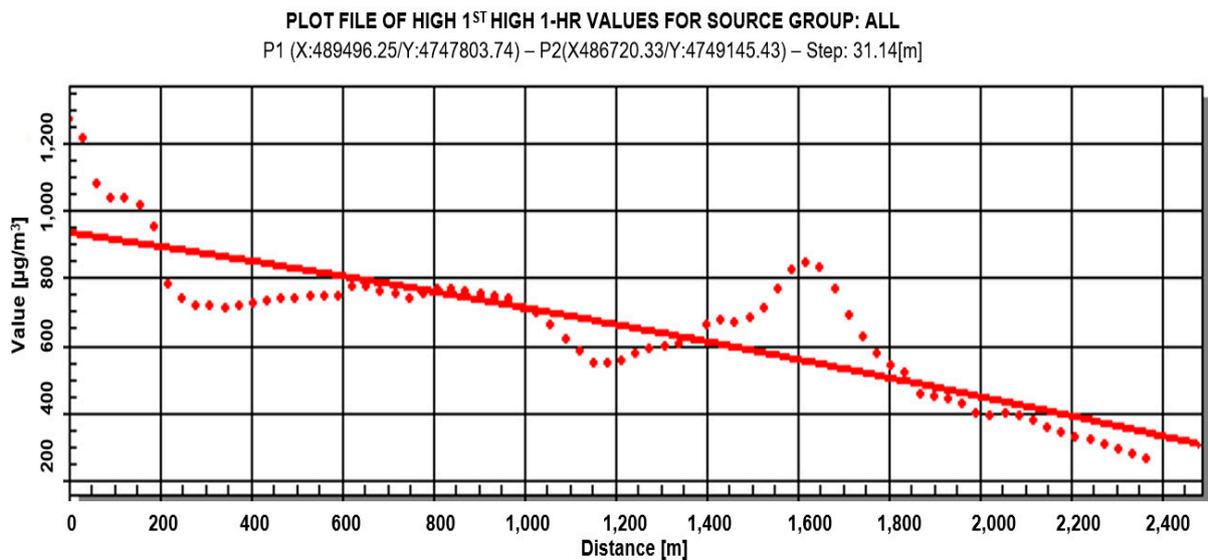


Figure 9. The highest concentration plot in the southwest direction.

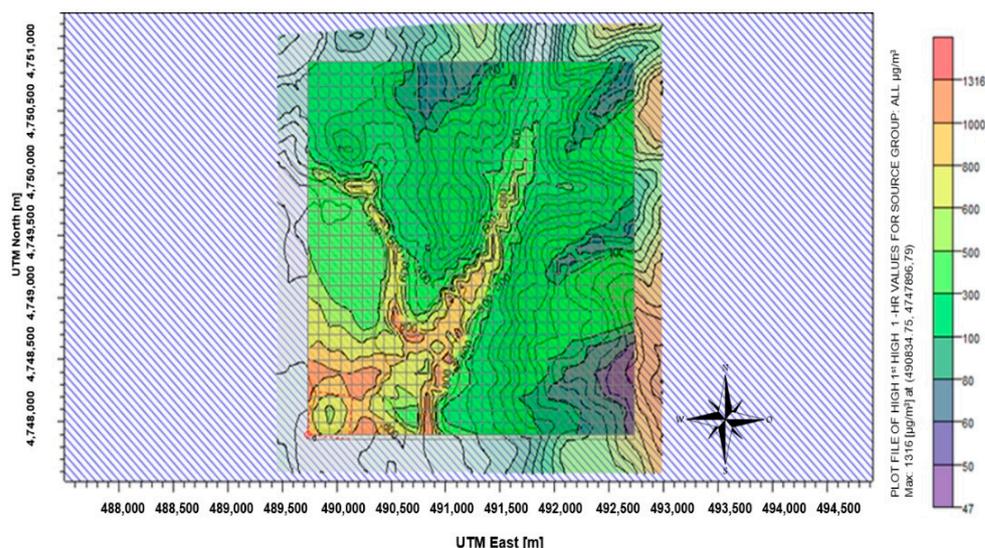


Figure 10. The highest calculated concentration for extreme wind conditions.

The weather conditions also changed in the case of relative humidity, so the emission of the solid particles was calculated for the case of minimum relative humidity of 14% and wind speed stronger than 11 m s^{-1} . The results of the PM 10 concentrations measured at the site for 2022 show good agreement with the calculations, as shown in Figure 11.

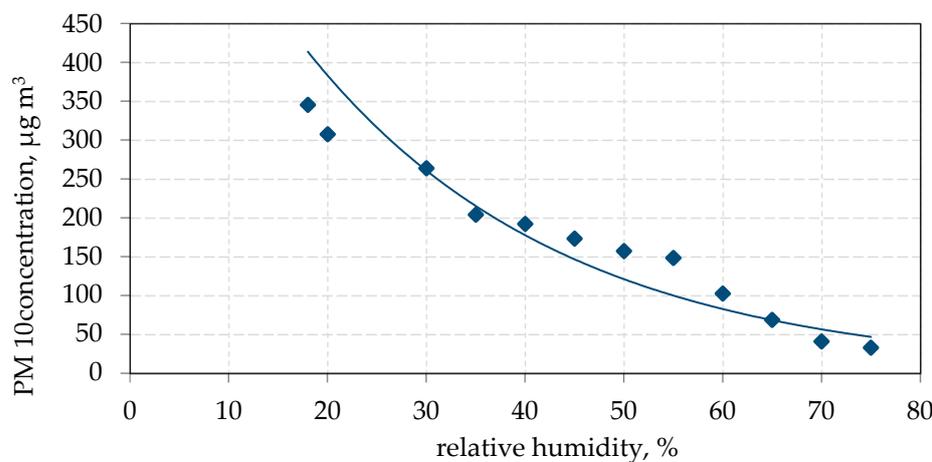


Figure 11. PM 10 concentration measured by the PM probe, collected and processed in relation to the relative humidity in August 2022.

The results of the PM 10 concentrations measured at the measuring point in the close vicinity of the deposit in the southwest direction in August 2022 show good agreement with the calculations for the given conditions. The PM 10 concentration of 324 µg m^{-3} was found to be the maximum for these conditions using AERMOD modeling, and 346 µg m^{-3} was the value obtained by measuring on-site. Comparing the EU threshold for PM 10 concentration, 40 µg m^{-3} , it can be concluded that the maximum 1-hour concentration exceeds the EU threshold by 9 times. The modeling shows the results presented in Figure 12 in the form of plume animation for the changing weather conditions. The plume animation was used to visualize the impact of the jarosite landfill on the city and population in all wind directions and speeds. Figure 12 presents the typical airborne pollution dispersion in the case of wind blowing in the southwest direction. The plume animation shows that the city will be endangered in the case of the wind blowing from the eastern mountains toward the river valley, as shown in Figure 12. The color scale shows the yellow, orange, and red areas exceeding the maximum values, which cover the most populated city areas.

The 1 h average concentration shows that starting from the deposit (dashed red square), the wind will carry high amounts of PM₁₀ particles to the southwest.

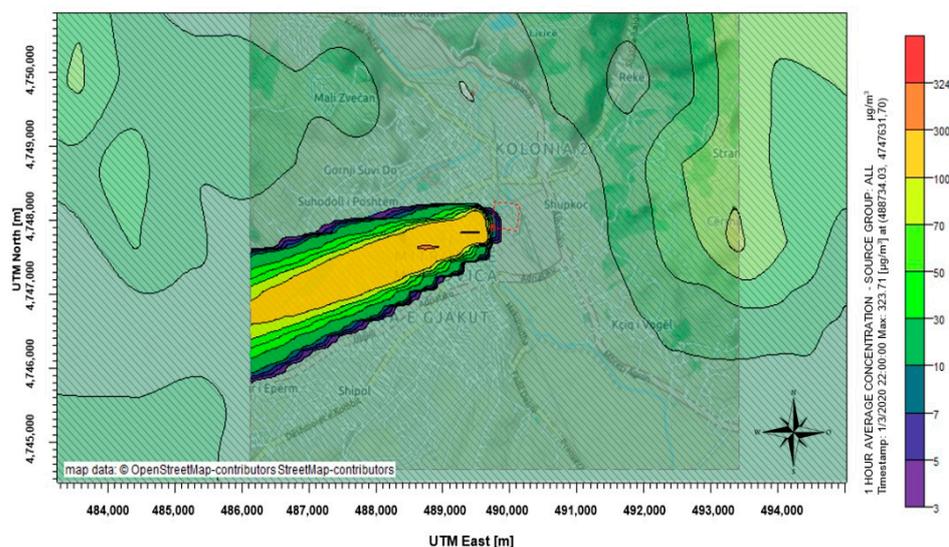


Figure 12. Plume animation cross-section of the pollution flow in the given weather conditions.

4. Conclusions

The jarosite waste deposit represents a constant and significant burden to the environment in the city of Mitrovica. Therefore, a comprehensive analysis of the jarosite waste deposit's actual and potential risks was presented to analyze the environmental impact and provide information for the decision on sustainable development planning for the city. The waste was analyzed for its toxicity, grain size, and mineral composition, and the range of airborne distribution was determined. The aspect of climate change was taken into consideration, as small-grained waste is scattered significantly in dry weather conditions. The results obtained using the AERMOD software package correspond to the measured values in average conditions, $71.5 \mu\text{g m}^{-3}$, and $324 \mu\text{g m}^{-3}$ for extremely dry and windy conditions recorded in August 2022. In addition, the direction west was detected as the most polluted in accordance with the terrain and wind rose for the location of the city. The usage of AERMOD is recommended for urban planning, but in the situation of climate change, relative humidity must be also taken into consideration. Future investigation should include downwash.

Author Contributions: Conceptualization, M.K.M., J.D., and L.Ć.; methodology, M.K.M.; software, J.D.; validation, M.K.M., J.D., and L.Ć.; formal analysis, M.K.M.; investigation, M.K.M.; resources, J.D., L.Ć.; data curation, M.K.M.; writing—original draft preparation M.K.M., J.D.; writing—review and editing, M.K.M., J.D., and L.Ć. All authors have read and agreed to the published version of the manuscript.

Funding: This research received no external funding.

Institutional Review Board Statement: Our work is original. It was not published in another journal, and it has not been submitted to another journal.

Informed Consent Statement: The authors consent to their participation in the article entitled “Modeling of atmospheric dispersion of jarosite particles from tailing waste in Mitrovica, Kosovo” written by Mihone Kerolli Mustafa, Jelena Djokic, and Lidija Ćurković.

Data Availability Statement: Not applicable.

Conflicts of Interest: The authors declare no conflict of interest.

References

1. Picazo-Rodríguez, N.G.; Carrillo-Pedroza, F.R.; Soria-Aguilar, M.D.J.; Baltierra, G.; González, G.; Martínez-Luevanos, A.; Almaguer Guzmán, I. Use of Thermally Modified Jarosite for the Removal of Hexavalent Chromium by Adsorption. *Crystals* **2022**, *12*, 80. [CrossRef]
2. Levei, E.; Frentiu, T.; Ponta, M.; Tanaselia, C.; Borodi, G. Characterization and assessment of potential environmental risk of tailings stored in seven impoundments in the Aries River basin, Western Romania. *Chem. Cent. J.* **2013**, *7*, 1–14. [CrossRef] [PubMed]
3. Chen, J.; Li, S.; Qu, Z.; Li, Z.; Wang, D.; Shen, J.; Li, Y. Study on Oxygen Evolution Reaction Performance of Jarosite/Composites. *Materials* **2022**, *15*, 668. [CrossRef] [PubMed]
4. Pappu, A.; Saxena, M.; Asolekar, S.R. Jarosite characteristics and its utilization potentials. *Sci. Total Environ.* **2005**, *359*, 232–243. [CrossRef]
5. Kerolli-Mustafa, M.; Fajković, H.; Rončević, S.; Ćurković, L. Assessment of metal risks from different depths of jarosite tailing waste of Trepča Zinc Industry, Kosovo based on BCR procedure. *J. Geochem. Explor.* **2015**, *148*, 161–168. [CrossRef]
6. Figueiredo, M.O.; da Silva, T.P. The Positive Environmental Contribution of Jarosite by Retaining Lead in Acid Mine Drainage Areas. *Int. J. Environ. Res. Public Health* **2011**, *8*, 1575–1582. [CrossRef]
7. Cruells, M.; Roca, A. Jarosites. Formation, Structure, Reactivity and Environmental. *Metals* **2022**, *12*, 802. [CrossRef]
8. Kerolli-Mustafa, M.; Ćurković, L.; Fajković, H.; Rončević, S. Ecological Risk Assessment of Jarosite Waste Disposal. *Croat. Chem. Acta* **2015**, *88*, 189–196. [CrossRef]
9. Lilic, N.; Cvjetic, A.; Knezevic, D.; Milisavljevic, V.; Pantelic, U. Dust and Noise Environmental Impact Assessment and Control in Serbian Mining Practice. *Minerals* **2018**, *8*, 34. [CrossRef]
10. Zhao, Z.; Xi, H.; Russo, A.; Du, H.; Gong, Y.; Xiang, J.; Zhou, Z.; Zhang, J.; Li, C.; Zhou, C. The Influence of Multi-Scale Atmospheric Circulation on Severe Haze Events in Autumn and Winter in Shanghai, China. *Sustainability* **2019**, *11*, 5979. [CrossRef]
11. Castillo, S.; de la Rosa, J.D.; Sánchez de la Campa, A.M.; González-Castanedo, Y.; Fernández-Caliani, J.C.; Gonzalez, I.; Romero, A. Contribution of mine wastes to atmospheric metal deposition in the surrounding area of an abandoned heavily polluted mining district (Rio Tinto mines, Spain). *Sci. Total Environ.* **2013**, *449*, 363–372. [CrossRef]
12. Pan, Y.; Wu, Z.; Zhou, J.; Zhao, J.; Ruan, X.; Liu, J.; Qian, G. Chemical characteristics and risk assessment of typical municipal solid waste interaction (MSWI) fly ash in China. *J. Hazard. Mater.* **2013**, *261*, 269–276. [CrossRef]
13. Nanni, A.; Tinarelli, G.; Solisio, C.; Pozzi, C. Comparison between Puff and Lagrangian Particle Dispersion Models at a Complex and Coastal Site. *Atmosphere* **2022**, *13*, 508. [CrossRef]
14. Leonor Maria Turtos Carbonell, L.M.; Gacita, M.S.; Rivero Oliva, J.; Garea, L.C.; Rivero, N.D.; Meneses Ruiz, E.M. Methodological guide for implementation of the AERMOD system with incomplete local data. *Atmos. Pollut. Res.* **2010**, *1*, 102–111. [CrossRef]
15. U.S. EPA. AERMOD: Description of Model Formulation. 2004. Available online: http://www.eng.utoledo.edu/aprg/courses/dm/aermod/aermod_mfd.pdf (accessed on 15 March 2022).
16. U.S. EPA. Technology Transfer Network Support Center for Regulatory Atmospheric Modelling. 2014. Available online: <https://www.epa.gov/scram/clean-air-act-permit-modeling-guidance> (accessed on 20 April 2022).
17. Zunaira, A.; Chen, Z.; Han, Y. Air quality modeling for effective environmental management in the mining region. *J. Air Waste Manag. Assoc.* **2018**, *68*, 1001–1014. [CrossRef]
18. Asif, Z.; Chen, Z. Environmental management in North American mining sector. *Environ. Sci. Pollut. Res.* **2016**, *23*, 167–179. [CrossRef]
19. Atabi, F.; Jafarigol, F.; Moattar, F.; Nouri, J. Comparison of AERMOD and CALPUFF models for simulating SO₂ concentrations in a gas refinery. *Environ. Monit. Assess.* **2016**, *188*, 516. [CrossRef]
20. Gobbi, G.P.; Liberto, L.D.; Barnaba, F. Impact of port emissions on EU-regulated and non-regulated air quality indicators: The case of Civitavecchia (Italy). *Sci. Total Environ.* **2020**, *719*, 134–984. [CrossRef]
21. European Environment Agency (EEA). The Application of Models under the European Union’s Air Quality Directive: A Technical Report. 2011. Available online: <https://www.eea.europa.eu/publications/fairmode> (accessed on 20 March 2022).
22. Thunis, P.; Pisoni, E.; Degraeuwe, B.; Kranenburg, R.; Schaap, M.; Clappier, A. Dynamic evaluation of air quality models over European regions. *Atmos. Environ.* **2015**, *111*, 185–194. [CrossRef]
23. Wang, Y.; Wu, X.; He, S. Eco-environmental assessment model of the mining area in Gongy. *China Inf. Sci.* **2021**, *11*, 17549. [CrossRef]
24. Holmes, N.S.; Morawska, L.A. Review of dispersion modeling and its application to the dispersion of particles: An overview of different dispersion models available. *Atmos. Environ.* **2006**, *40*, 5902–5928. [CrossRef]